

# Ecological site R025XY051NV ERODED CLAYPAN 12-16 P.Z.

Last updated: 4/25/2024 Accessed: 08/17/2024

### **General information**

**Provisional**. A provisional ecological site description has undergone quality control and quality assurance review. It contains a working state and transition model and enough information to identify the ecological site.

#### **MLRA** notes

Major Land Resource Area (MLRA): 025X–Owyhee High Plateau

#### MLRA Notes 25—Owyhee High Plateau

This area is in Nevada (56 percent), Idaho (30 percent), Oregon (12 percent), and Utah (2 percent). It makes up about 27,443 square miles. MLRA 25 is characteristically cooler and wetter than the neighboring MLRAs of the Great Basin. The western boundary is marked by a gradual transition to the lower and warmer basins of MLRA 24. The boundary to the south-southeast, with MLRA 28B, is marked by gradual changes in geology marked by an increased dominance of singleleaf pinyon and Utah juniper and a reduced presence of Idaho fescue. The boundary to the north, with MLRA 11, is a rapid transition from the lava plateau topography to the lower elevation Snake River Plain.

Physiography:

All of this area lies within the Intermontane Plateaus. The southern half is in the Great Basin section of the Basin and Range province. This part of the MLRA is characterized by isolated, uplifted fault-block mountain ranges separated by narrow, aggraded desert plains. This geologically older terrain has been dissected by numerous streams draining to the Humboldt River.

The northern half of the area lies within the Columbia Plateaus province. This part of the MLRA forms the southern boundary of the extensive Columbia Plateau basalt flows. Most of the northern half is in the Payette section, but the northeast corner is in the Snake River Plain section. Deep, narrow canyons draining into the Snake River have been incised into this broad basalt plain. Elevation ranges from 3,000 to 7,550 feet on rolling plateaus and in gently sloping basins. It is more than 9,840 feet on some steep mountains. The Humboldt River crosses the southern half of this area

Geology:

The dominant rock types in this MLRA are volcanic. They include andesite, basalt, tuff, and rhyolite. In the north and west parts of the area, Cretaceous granitic rocks are exposed among Miocene volcanic rocks in mountains. A Mesozoic igneous and metamorphic rock complex dominates the south and east parts of the area. Upper and Lower Paleozoic calcareous sediments, including oceanic deposits, are exposed with limited extent in the mountains. Alluvial fan and basin fill sediments occur in the valleys.

Climate:

The average annual precipitation in most of this area is typically 11 to 22 inches. It increases to as much as 49 inches at the higher elevations. Rainfall occurs in spring and sporadically in summer. Precipitation occurs mainly as snow in winter. The precipitation is distributed fairly evenly throughout fall, winter, and spring. The amount of precipitation is lowest from midsummer to early autumn. The average annual temperature is 33 to 51 degrees F. The freeze-free period averages 130 days and ranges from 65 to 190 days, decreasing in length with elevation. It is typically less than 70 days in the mountains. Water:

The supply of water from precipitation and streamflow is small and unreliable, except along the Owyhee, Bruneau, and Humboldt Rivers. Streamflow depends largely on accumulated snow in the mountains. Surface water from mountain runoff is generally of excellent quality and suitable for all uses. The basin fill sediments in the narrow alluvial valleys between the mountain ranges provide some ground water for irrigation. The alluvial deposits along the large streams have the most ground water. Based on measurements of water quality in similar deposits in

adjacent areas, the basin fill deposits probably contain moderately hard water. The water is suitable for almost all uses. The carbonate rocks in this area are considered aquifers, but they are little used. Springs are common along the edges of the limestone outcrops. Soils:

The dominant soil orders in this MLRA are Aridisols and Mollisols. The soils in the area dominantly have a mesic or frigid temperature regime and an aridic, aridic bordering on xeric, or xeric moisture regime. Soils with aquic moisture regimes are limited to drainage or spring areas, where moisture originates or runs on and through. These soils are of a very limited extent throughout the MLRA. They generally are well drained, clayey or loamy, and shallow or moderately deep. Most of the soils formed in mixed parent material. Volcanic ash and loess mantle the landscape. Surface soil textures are loam and silt loam with ashy texture modifiers in some areas. Argillic horizons occur on the more stable landforms. They are exposed nearer the soil surface on convex landforms, where ash and loess deposits are more likely to erode. Soils that formed in carbonatic parent material in areas that receive less than 12 inches of precipitation are characterized by calcic horizons in the upper part of the profile. Soils that formed on stable landforms at the lower elevations are dominated by ochric horizons. Soils that formed at the middle and upper elevations are characterized by mollic epipedons. Soils in drainage areas at all elevations that receive moisture running on or through them are characterized by thicker mollic epipedons. Biological Resources:

This MLRA supports shrub-grass vegetation. Lower elevations are characterized by Wyoming big sagebrush associated with bluebunch wheatgrass, western wheatgrass, and Thurber's needlegrass. Other important plants include bluegrass, squirreltail, penstemon, phlox, milkvetch, lupine, Indian paintbrush, aster, and rabbitbrush. Black sagebrush occurs but is less extensive. Singleleaf pinyon and Utah juniper occur in limited areas. With increasing elevation and precipitation, vast areas characterized by mountain big sagebrush or low sagebrush/early sagebrush in association with Idaho fescue, bluebunch wheatgrass, needlegrasses, and bluegrass become common. Snowberry, curl-leaf mountain mahogany, ceanothus, and juniper also occur. Mountains at the highest elevations support whitebark pine, Douglas-fir, limber pine, Engelmann spruce, subalpine fir, aspen, and curl-leaf mountain mahogany.

Major wildlife species include mule deer, bighorn sheep, pronghorn, mountain lion, coyote, bobcat, badger, river otter, mink, weasel, golden eagle, red-tailed hawk, ferruginous hawk, Swainson's hawk, northern harrier, prairie falcon, kestrel, great horned owl, short-eared owl, long-eared owl, burrowing owl, pheasant, sage grouse, chukar, gray partridge, and California quail. Reptiles and amphibians include western racer, gopher snake, western rattlesnake, side-blotched lizard, western toad, and spotted frog. Fish species include bull, red band, and rainbow trout.

## **Ecological site concept**

This site occurs on sideslopes of mountains, hills, erosional fan remnants and rock-pediment remnants on all aspects. Slopes range from 2 to 75 percent, but slope gradients of 4 to 30 percent are most typical. Elevations are 6000 to 7000 feet.

The soils associated with this site are very shallow and shallow and well drained. The soil surface has high percentages of gravels and/or cobbles. Depth to bedrock ranges from 5 to 12 inches. Runoff is medium and permeability is moderately slow. Potential for sheet and rill erosion is moderate to high and soils may reflect a past erosion history in a truncated A horizon or in thin surface soils.

The reference plant community is dominated by Idaho fescue, bluebunch wheatgrass and low sagebrush. Shrubs dominate the aspect of this site. Antelope bitterbrush, Thurber's needlegrass and Sandberg's bluegrass are other important species associated with this site.

## **Associated sites**

R025XY012NV	LOAMY SLOPE 12-16 P.Z.
R025XY017NV	CLAYPAN 12-16 P.Z.

## Similar sites

R025XY024NV	<b>MOUNTAIN RIDGE</b> FEID-POA dominant grasses; occurs on mountain summits and upper backslopes
R025XY022NV	COBBLY CLAYPAN 8-12 P.Z. PSSPS-ACTH7 codominant grasses
R025XY017NV	CLAYPAN 12-16 P.Z. More productive site
R025XY018NV	CLAYPAN 10-12 P.Z. FEID minor to rare species; more productive site
R025XY044ID	VERY SHALLOW STONY LOAM 10-14

#### Table 1. Dominant plant species

Tree	Not specified
Shrub	(1) Artemisia arbuscula
Herbaceous	<ul><li>(1) Festuca idahoensis</li><li>(2) Pseudoroegneria spicata</li></ul>

### **Physiographic features**

This site occurs on sideslopes of mountains, hills, erosional fan remnants and rock-pediment remnants on all aspects. Slopes range from 2 to 75 percent, but slope gradients of 4 to 30 percent are most typical. Elevations are 6000 to 7000 feet.

#### Table 2. Representative physiographic features

Landforms	<ul><li>(1) Mountain slope</li><li>(2) Hill</li><li>(3) Erosion remnant</li></ul>
Runoff class	High to very high
Flooding frequency	None
Ponding frequency	None
Elevation	1,829–2,134 m
Slope	2–75%
Water table depth	183 cm
Aspect	W, NW, N, NE, E, SE, S, SW

### **Climatic features**

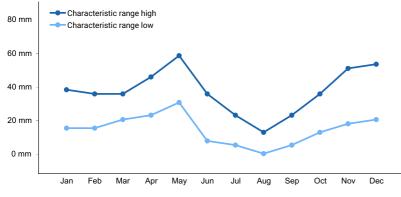
The climate associated with this site is semiarid, characterized by cold, moist winters and warm, dry summers. The average annual precipitation ranges from 12 to 14 inches. Mean annual air temperature is typically <45 degrees F.

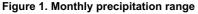
Mean annual precipitation across the range in which this ES occurs is 13.13".

Monthly mean precipitation: January 1.45"; February 1.12"; March 1.20"; April 1.21"; May 1.62"; June 1.07"; July 0.55"; August 0.49"; September 0.62"; October 0.83"; November 1.41"; December 1.54".

\*The above data is averaged from the Mountain City RS and Wild horse RSVR WRCC climate stations.

Frost-free period (average)	23 days
Freeze-free period (average)	59 days
Precipitation total (average)	330 mm





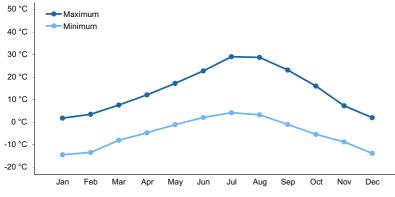


Figure 2. Monthly average minimum and maximum temperature

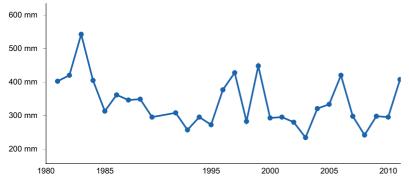


Figure 3. Annual precipitation pattern

### **Climate stations used**

- (1) MTN CITY RS [USC00265392], Mountain City, NV
- (2) WILD HORSE RSVR [USC00269072], Deeth, NV

#### Influencing water features

There are no influencing water features associated with this site.

### **Soil features**

The soils associated with this site are very shallow to shallow and well drained. The soil surface has high percentages of gravels and/or cobbles. Depth to bedrock ranges from 5 to 12 inches. Runoff is medium and

permeability is moderately slow. Potential for sheet and rill erosion is moderate to high and soils may reflect a past erosion history in a truncated A horizon or in thin surface soils.

The soil series correlated with this site include: Bregar.

The representative soil series is Bregar, classified as a loamy-skeletal, mixed, superactive, frigid Lithic Xeric Haplargid. This soil is very shallow to shallow, well drained, and formed in residuum and colluvium derived from andesite, tuff, and quartzite. Reaction is slightly acid through slightly alkaline. Diagnostic horizons include an ochric epidedon that occurs from the soil surface to 6 inches and an argillic horizon that occurs from 6 inches to 11 inches. Clay content in the particle-size control section averages 18 to 30 percent. Rock framents average 35 to 70 percent, mainly gravel and cobbles. Lithology of fragments is mainly volcanic rocks such as andesite.

Parent material	<ul><li>(1) Colluvium</li><li>(2) Residuum</li><li>(3) Alluvium</li></ul>
Surface texture	<ul><li>(1) Very gravelly sandy loam</li><li>(2) Very cobbly loam</li></ul>
Family particle size	(1) Loamy-skeletal
Drainage class	Well drained
Permeability class	Moderately rapid
Depth to restrictive layer	13–30 cm
Soil depth	13–30 cm
Surface fragment cover <=3"	20–25%
Surface fragment cover >3"	45–50%
Available water capacity (0-101.6cm)	1.52–2.29 cm
Calcium carbonate equivalent (0-101.6cm)	0%
Electrical conductivity (0-101.6cm)	0 mmhos/cm
Sodium adsorption ratio (0-101.6cm)	0
Soil reaction (1:1 water) (0-101.6cm)	6.1–7.8
Subsurface fragment volume <=3" (Depth not specified)	20–34%
Subsurface fragment volume >3" (Depth not specified)	2–23%

#### Table 4. Representative soil features

## **Ecological dynamics**

An ecological site is the product of all the environmental factors responsible for its development and has a set of key characteristics that influence a site's resilience to disturbance and resistance to invasives. Key characteristics include 1) climate (precipitation and temperature), 2) topography (aspect, slope, elevation, and landform), 3) hydrology (infiltration and runoff), 4) soils (depth, texture, structure and organic matter), 5) plant communities (functional groups and productivity), and 6) natural disturbance regime (fire, herbivory, etc.) (Caudle 2013). Biotic factors that influence resilience include site productivity, species composition and structure, and population regulation and regeneration (Chambers et al. 2013).

This ecological site is dominated by deep-rooted cool season, perennial bunchgrasses and long-lived shrubs (50+ years) with high root to shoot ratios. Community types with low sagebrush as the dominant shrub, however, were

found to have soil depths and thus available rooting depths of 71 to 81 centimeters in a study in northeast Nevada (Jensen 1990). These shrubs have a flexible generalized root system with development of both deep taproots and laterals near the surface (Comstock and Ehleringer 1992).

Periodic drought regularly influences sagebrush ecosystems and drought duration and severity has increased throughout the 20th century in much of the Intermountain West. Major shifts away from historical precipitation patterns have the greatest potential to alter ecosystem function and productivity. Species composition and productivity can be altered by the timing of precipitation and water availability with the soil profile (Bates et al. 2006).

Native insect outbreaks are also important drivers of ecosystem dynamics in sagebrush communities. Climate is generally believed to influence the timing of insect outbreaks, especially with regard to sagebrush defoliator Aroga moth (Aroga websteri). Aroga moth infestations have occurred throughout the Great Basin in the 1960s, early 1970s, and has been ongoing in Nevada since 2004 (Bentz et al. 2008). Thousands of acres of big sagebrush have been impacted, with partial to complete die-off observed (Gates 1964, Hall 1965); the research is inconclusive regarding the damage sustained by black sagebrush populations.

Low sagebrush is fairly drought tolerant but also tolerates periodic wetness during a portion of the growing season. Low sagebrush is also susceptible to Aroga moth, which can partially or entirely kill individual plants or entire stands of big sagebrush (Furniss and Barr 1975), though the research is inconclusive of the damage sustained by low sagebrush populations.

The Great Basin sagebrush communities have high spatial and temporal variability in precipitation both among years and within growing seasons. Nutrient availability is typically low but increases with elevation and closely follows moisture availability. The invasibility of plant communities is often linked to resource availability. Disturbance can decrease resource uptake due to damage or mortality of the native species and depressed competition. It can also increase resource pools by the decomposition of dead plant material following disturbance. The invasion of sagebrush communities by cheatgrass (*Bromus tectorum*) has been linked to disturbances (fire, abusive grazing) that have resulted in fluctuations in resources (Chambers et al. 2007).

The perennial bunchgrasses that are dominant on this ecological site includes Idaho fescue and bluebunch wheatgrass. Perennial bunchgrasses generally have somewhat shallower root systems than shrubs in these systems, but root densities are often as high as or higher than those of shrubs in the upper 0.5 m but taper off more rapidly than shrubs. General differences in root depth distributions between grasses and shrubs result in resource partitioning in these shrub/grass systems.

As ecological condition delines, the dwarf sagebrushes and small rabbitbrush become dominant with increases of Sandberg's bluegrass, phlox and other mat forming forbs in the understory. The potential invasive/noxious weeds are cheatgrass, rabbitbrush and snakeweed.

This ecological site has low to moderate resilience to disturbance and resistance to invasion. Increased resilience increases with elevation, aspect, increased precipitation and increased nutrient availability. One possible alternative stable state has been identified for this ecological site.

### Fire Ecology:

Prior to 1897, mean fire return intervals for low sagebrush communities have been estimated to be from 35 to over 100 years. Fire most often occurs during wet years with high forage production.

Low sagebrush is killed by fire and does not sprout (Tisdale and Hironaka 1984). Establishment after fire is from seed, generally blown in and not from the seed bank (Bradley et al. 1992). Fire risk is greatest following a wet, productive year when there is greater production of fine fuels (Beardall and Sylvester 1976). Recovery time of low sagebrush following fire is variable (Young 1983). After fire, if regeneration conditions are favorable, low sagebrush recovers in 2 to 5 years; on harsh sites where cover is low to begin with and/or erosion occurs after fire, recovery may require more than 10 years (Young 1983). Slow regeneration may subsequently worsen erosion (Blaisdell et al. 1982).

Antelope bitterbrush, a minor component on this site, is moderately fire tolerant (McConnell and Smith 1977). It regenerates by seed and resprouting (Blaisdell and Mueggler 1956, McArthur et al. 1982), however sprouting ability is highly variable and has been attributed to genetics, plant age, phenology, soil moisture and texture and fire

severity (Blaisdell and Mueggler 1956, Blaisdell et al. 1982, Clark et al. 1982, Cook et al. 1994). Bitterbrush sprouts from a region on the stem approximately 1.5 inches above and below the soil surface; the plant rarely sprouts if the root crown is killed by fire (Blaisdell and Mueggler 1956). Low intensity fires may allow for bitterbrush to sprout; however, community response also depends on soil moisture levels at time of fire (Murray 1983). Lower soil moisture allows more charring of the stem below ground level (Blaisdell and Mueggler 1956), thus sprouting will usually be more successful after a spring fire than after a fire in summer or fall (Murray 1983, Busse et al. 2000, Kerns et al. 2006). If cheatgrass is present, bitterbrush seedling success is much lower. The factor that most limits establishment of bitterbrush seedlings is competition for water resources with cheatgrass, an invasive species (Clements and Young 2002).

The effect of fire on bunchgrasses relates to culm density, culm-leaf morphology, and the size of the plant. The initial condition of bunchgrasses within the site along with seasonality and intensity of the fire all factor into the individual species response. The growing points for most forbs and grasses are located at or below the soil surface, providing relative protection from disturbances which decrease above ground biomass, such as grazing or fire. Thus, fire mortality is more correlated to duration and intensity of heat which is related to culm density, culm-leaf morphology, size of plant and abundance of old growth (Wright 1971, Young 1983). However, season and severity of the fire and post-fire soil moisture

availability will influence plant response.

Idaho fescue grows in a dense, fine-leaved tuft. Fires tend to burn within the accumulated fine leaves at the base of the plant and may produce temperatures sufficient to kill some of the root crown. Mature Idaho fescue plants are commonly reported to be severely damaged by fire in all seasons.

Bluegrass is generally unharmed by fire. It produces little litter, and its small bunch size and sparse litter reduces the amount of heat transferred to perennating buds in the soil. Its rapid maturation in the spring also reduces fire damage, since it is dormant when most fires occur. Sandberg bluegrass has been found to increase following fire, likely due to its low stature and productivity (Daubenmire 1975) and may retard reestablishment of deeper rooted bunchgrasses.

Bluebunch wheatgrass, a minor component of this ecological site, has coarse stems with little leafy material, the aboveground biomass burns rapidly and little heat is transferred downward into the crowns (Young 1983). Bluebunch wheatgrass was described as fairly tolerant of burning, other than in May in eastern Oregon (Britton et al. 1990). Uresk et al. (1976) reported burning increased vegetative and reproductive vigor of bluebunch wheatgrass, thus it experiences slight damage to fire but is more susceptible in drought years (Young 1983).

Sandberg bluegrass has been found to increase following fire, likely due to its low stature and productivity (Daubenmire 1975) and may retard reestablishment of deeper rooted bunchgrasses.

### State and transition model

#### MLRA 25 Eroded Claypan 12-16" 025XY051NV

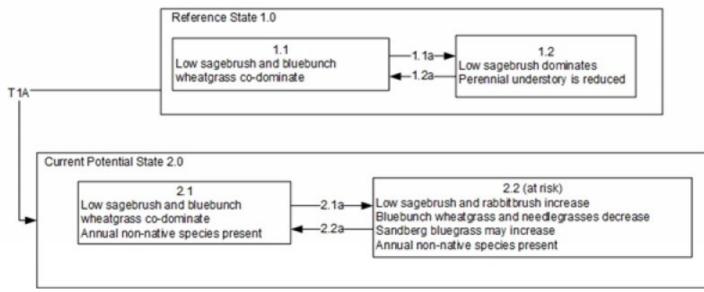


Figure 5. T. Stringham July 2015

#### MLRA 25 Eroded Claypan 12-16" 025XY051NV Legend

Reference State 1.0 Community Phase Pathways 1.1a: Long term drought and/or herbivory may reduce perennial bunchgrasses 1.2a: Release from drought and/or herbivory

Transition T1A: Introduction of non-native annual species.

Current Potential State 2.0 Community Phase Pathways 2.1a: Long term drought and/or inappropriate grazing may reduce perennial bunchgrasses 2.2a: Release from drought and/or grazing management

Figure 6. T. Stringham July 2015

### State 1 Reference State

The Reference State 1.0 is a representative of the natural range of variability under pristine conditions. The reference state has three general community phases: a shrub-grass dominant phase, a perennial grass dominant phase and a shrub dominant phase. State dynamics are maintained by interactions between climatic patterns and disturbance regimes. Negative feedbacks enhance ecosystem resilience and contribute to the stability of the state. These include the presence of all structural and functional groups, low fine fuel loads, and retention of organic matter and nutrients. Plant community phase changes are primarily driven by fire, periodic drought and/or insect or disease attack.

### Community 1.1 Low sagebrush-bluebunch wheatgrass

The reference plant community is dominated by Idaho fescue, bluebunch wheatgrass and low sagebrush. Shrubs dominate the aspect of this site. Antelope bitterbrush, Thurber's needlegrass and Sandberg's bluegrass are other important species associated with this site. Potential vegetative composition is about 40 percent grasses, 10 percent forbs and 50 percent shrubs. Approximate ground cover (basal and crown) is 15 to 25 percent.

Table 5. Annual production by plant type

Plant Type	Low (Kg/Hectare)	Representative Value (Kg/Hectare)	
Shrub/Vine	112	168	280
Grass/Grasslike	90	135	224
Forb	22	34	56
Total	224	337	560

## Community 1.2 Low sagebrush

This community phase is characteristic of a post-disturbance, early/mid-seral community. Low sagebrush dominates the overstory and deep-rooted perennial bunchgrasses in the understory are reduced.

## Pathway 1.1a Community 1.1 to 1.2

Long term drought and/or herbivory may reduce perennial bunchgrasses.

### Pathway 1.2a Community 1.2 to 1.1

Time and lack of disturbance will allow perennial bunchgrasses to increase.

## State 2 Current Potential State

This state is similar to the Reference State 1.0 and has two community phases. Ecological function has not changed, however the resiliency of the state has been reduced by the presence of invasive weeds. These nonnative species can be highly flammable, and promote fire where historically fire had been infrequent. Negative feedbacks enhance ecosystem resilience and contribute to the stability of the state. These feedbacks include the presence of all structural and functional groups, low fine fuel loads, and retention of organic matter and nutrients. Positive feedbacks decrease ecosystem resilience and stability of the state. These include the non-natives' high seed output, persistent seed bank, rapid growth rate, ability to cross pollinate, and adaptations for seed dispersal.

## Community 2.1 Low sagebrush/bluebunch wheatgrass/annual non-native species present

This community phase is similar to the Reference State Community Phase 1.1, with the presence of non-native species in trace amounts. Sagebrush, bluebunch wheatgrass and Thurber's needlegrass dominate the site. Forbs and other shrubs and grasses make up smaller components of this site.

## Community 2.2 Low sagebrush/bluegrass/annual non-native species present (at risk)

This community is at risk of crossing a threshold to another state. Sagebrush dominates the overstory and perennial bunchgrasses in the understory are reduced, either from competition with shrubs or from inappropriate grazing management, or from both. Rabbitbrush may be a significant component. Sandberg bluegrass may increase and become co-dominate with deep rooted bunchgrasses. Annual non-natives species may be stable or increasing due to lack of competition with perennial bunchgrasses. This site is susceptible to further degradation from inappropriate grazing management, drought, and fire.

Pathway 2.1a Community 2.1 to 2.2 Long term drought and/or inappropriate grazing may reduce perennial bunchgrasses.

## Pathway 2.2a Community 2.2 to 2.1

Time and lack of disturbance and/or grazing management that favors the establishment and growth of sagebrush allows the shrub component to recover. The establishment of low sagebrush can take many years.

## Transition T1A State 1 to 2

Trigger: This transition is caused by the introduction of non-native annual plants, such as cheatgrass, mustards, and bur buttercup. Slow variables: Over time the annual non-native species will increase within the community. Threshold: Any amount of introduced non-native species causes an immediate decrease in the resilience of the site. Annual non-native species cannot be easily removed from the system and have the potential to significantly alter disturbance regimes from their historic range of variation.

## Additional community tables

 Table 6. Community 1.1 plant community composition

Group	Common Name	Symbol	Scientific Name	Annual Production (Kg/Hectare)	Foliar Cover (%)
Grass	/Grasslike		· · · · ·	·	
1	Primary Perennial Gr	Primary Perennial Grasses			
	Idaho fescue	FEID	Festuca idahoensis	50–84	-
	bluebunch wheatgrass	PSSPS	Pseudoroegneria spicata ssp. spicata	34–67	_
	Sandberg bluegrass	POSE	Poa secunda	7–27	_
	Thurber's needlegrass	ACTH7	Achnatherum thurberianum	7–27	_
2	Secondary Perennial	Grasses		7–17	
	Indian ricegrass	ACHY	Achnatherum hymenoides	2–7	_
	western needlegrass	ACOC3	Achnatherum occidentale	2–7	_
	squirreltail	ELEL5	Elymus elymoides	2–7	_
	basin wildrye	LECI4	Leymus cinereus	2–7	_
Forb			••	•	
3	Perennial Forbs			17–50	
	buckwheat	ERIOG	Eriogonum	2–10	-
	spiny phlox	PHHO	Phlox hoodii	2–10	-
	goldenweed	PYRRO	Pyrrocoma	2–10	_
Shrub	Vine				
4	Primary Shrubs			108–168	
	little sagebrush	ARAR8	Artemisia arbuscula	84–118	-
	antelope bitterbrush	PUTR2	Purshia tridentata	17–34	_
	yellow rabbitbrush	CHVI8	Chrysothamnus viscidiflorus	7–17	
5	Secondary Shrubs			7–17	
	rubber rabbitbrush	ERNA10	Ericameria nauseosa	2–7	_

### **Animal community**

Livestock Interpretations:

This site is suited to livestock grazing. Considerations for grazing management include timing, intensity and duration of grazing.

In general, bunchgrasses best tolerate light grazing after seed formation. Britton et al. (1990) observed the effects of clipping date on basal area of 5 bunchgrasses in eastern Oregon, and found that grazing from August to October (after seed set) has the least impact. Heavy grazing during the growing season will reduce perennial bunchgrasses and increase sagebrush (Laycock 1967). Abusive grazing by cattle or horses will likely increase low sagebrush, rabbitbrush and some forbs such as arrowleaf balsamroot. Annual non-native weedy species such as cheatgrass and mustards, and potentially medusahead, may invade.

Reduced bunchgrass vigor or density provides an opportunity for Sandberg bluegrass expansion and/or cheatgrass and other invasive species to occupy interspaces. Bluegrass is a widespread, palatable forage grass that is one of the earliest grasses in the spring and is sought by domestic livestock and several wildlife species. Its production is closely tied to weather conditions; little forage is produced in drought years, making it a less dependable food source than other perennial bunchgrasses. Sandberg bluegrass increases under grazing pressure (Tisdale and Hironaka 1981) and is capable of co-existing with cheatgrass or other weedy species. Excessive sheep grazing favors Sandberg bluegrass; however, where cattle are the dominant grazers, cheatgrass often dominates (Daubenmire 1970). Thus, depending on the season of use, the grazer and site conditions, either Sandberg bluegrass or cheatgrass may become the dominant understory with inappropriate grazing management.

Bluebunch wheatgrass is moderately grazing-tolerant and is very sensitive to defoliation during the active growth period (Blaisdell and Pechanec 1949, Laycock 1967, Anderson and Scherzinger 1975, Britton et al. 1990). Herbage and flower stalk production was reduced with clipping at all times during the growing season; however, clipping was most harmful during the boot stage (Blaisdell and Pechanec 1949). Tiller production and growth of bluebunch was also greatly reduced when clipping was coupled with drought (Busso and Richards 1995). Mueggler (1975) estimated that low-vigor bluebunch wheatgrass may need up to 8 years rest to recover. Although an important forage species, it is not always the preferred species by livestock and wildlife.

Thurber's needlegrass is an important forage source for livestock and wildlife in the arid regions of the west (Ganskopp 1988). Although the seeds are not injurious, grazing animals avoid them when they begin to mature. Sheep, however, have been observed to graze the leaves closely, leaving stems untouched (Eckert and Spencer 1987). Heavy grazing during the growing season has been shown to reduce the basal area of Thurber's needlegrass (Eckert and Spencer 1987), suggesting that both seasonality and utilization are important factors in management of this plant. A single defoliation, particularly during the boot stage, was found to reduce herbage production and root mass thus potentially lowering the competitive ability of this needlegrass (Ganskopp 1988).

Domestic sheep and, to a much lesser degree, cattle consume low sagebrush, particularly during the spring, fall, and winter (Sheehy and Winward 1981). Heavy dormant season grazing by sheep will reduce sagebrush cover and increase grass production (Laycock 1967). Severe trampling damage to supersaturated soils may occur if sites are used in early spring when there is abundant snowmelt. Trampling damage, particularly from cattle or horses, in low sagebrush habitat types is greatest when high clay content soils are wet. In drier areas that contain more gravelly soils, no serious trampling damage occurs, even when the soils are wet (Hironaka et al. 1983).

Antelope bitterbrush is often utilized heavily by domestic livestock (Wood 1995). Grazing tolerance is dependent on site conditions (Garrison 1953) and the shrub can be severely hedged during the dormant season.

Stocking rates vary over time depending upon season of use, climate variations, site, and previous and current management goals. A safe starting stocking rate is an estimated stocking rate that is fine-tuned by the client by adaptive management through the year and from year to year.

### Wildlife Interpretations:

Low sagebrush is considered a valuable browse plant for wildlife during the spring, fall and winter months. In some areas, it is of little value in winter due to heavy snow. Mule deer utilize and sometimes prefer low sagebrush, particularly in winter and early spring.

Sagebrush-grassland communities provide critical sage-grouse breeding and nesting habitats. Open Wyoming

sagebrush communities are preferred nesting habitat. Meadows surrounded by sagebrush may be used as feeding and strutting grounds. Sagebrush is a crucial component of their diet year-round, and sage-grouse select sagebrush almost exclusively for cover. Leks are often located on low sagebrush sites, grassy openings, dry meadows, ridgetops, and disturbed sites.

Antelope bitterbrush is a critical browse species for mule deer, antelope and elk.

### **Recreational uses**

Aesthetic value is derived from the diverse floral and faunal composition and the colorful flowering of wild flowers and shrubs during the spring and early summer. This site offers rewarding opportunities to photographers and for nature study. This site is used for camping and hiking and has potential for upland and big game hunting.

### **Other information**

Low sagebrush can be successfully transplanted or seeded in restoration.

### Inventory data references

Soils and Physiographic features were gathered from NASIS.

### **Type locality**

Location 1: Elko County, NV		
Township/Range/Section	T38N R53E S22	
	About 4 air miles west northwest of the Stampede Ranch, and 4 miles south of NvHwy 11 (Taylor Canyon area), Independence Mountains, Elko County, Nevada.	

### **Other references**

Anderson, E. W. and R. J. Scherzinger. 1975. Improving quality of winter forage for elk by cattle grazing. Journal of Range Management 28: 120-125.

Baker, W. L. 2006. Fire and restoration of sagebrush ecosystems. Wildlife Society Bulletin 34: 177-185.

Barnett, J. K. and J. A. Crawford. 1994. Pre-laying nutrition of sage grouse hens in Oregon. Journal of Range Management 47: 114-118.

Barney, M. A. and N. C. Frischknecht. 1974. Vegetation changes following fire in the pinyon-juniper type of west-central Utah. Journal of Range Management 27: 91-96.

Bates, J. D., T. Svejcar, R. F. Miller, and R. A. Angell. 2006. The effects of precipitation timing on sagebrush steppe vegetation. Journal of Arid Environments 64: 670-697.

Beardall, L. E. and V. E. Sylvester. 1976. Spring burning for removal of sagebrush competition in Nevada. In: Tall Timbers Fire Ecology Conference and Proceedings. Tall Timbers Research Station. 14: 539-547

Blaisdell, J. P., R. B. Murray, and E. D. McArthur. 1982. Managing intermountain rangelands - sagebrush-grass ranges. Gen. Tech. Rep. INT-134. U.S. Department of Agriculture, Forest Service, Intermountain Forest and Range Experiment Station, Ogden, UT. p. 41.

Blaisdell, J. P. and W. F. Mueggler. 1956. Sprouting of bitterbrush (Purshia tridentata) following burning or top removal. Ecology 37:365-370.

Blaisdell, J. P. and J. F. Pechanec. 1949. Effects of herbage removal at various dates on vigor of bluebunch wheatgrass and arrowleaf balsamroot. Ecology 30:298-305.

Bradley, A. F., N. V. Noste, and W. C. Fischer. 1992. Fire ecology of forests and woodlands in Utah. Gen. Tech. Rep. INT-287. U.S. Department of Agriculture, Forest Service, Intermountain Research Station. P. 128.

Britton, C.M., F.A. Sneva, and R.G. Clark. 1979. Effect of harvest date on five bunchgrasses of eastern Oregon. In: 1979 Progress Report: Research in Rangeland Management. Special Report 549. Corvallis, OR: Oregon State University, Agricultural Experiment Station: Pgs 16-19. In cooperation with U.S. Department of Agriculture, SEA-AR.

Britton, C. M., G. R. McPherson, and F. A. Sneva. 1990. Effects of burning and clipping on five bunchgrasses in eastern Oregon. Great Basin Naturalist 50:115-120.

Bunting, S. 1994. Effects of fire on juniper woodland ecosystems in the Great Basin. In: S. Monsen, S. Kitchen [eds] Proceedings--Ecology and Management of Annual Rangelands Gen. Tech. Rep. INT-GTR-313. U.S. Department of Agriculture, Forest Service, Intermountain Research Station, Ogden, UT. p. 53-55

Busse, D., A. Simon, and M. Riegel. 2000. Tree-growth and understory responses to low-severity prescribed burning in thinned Pinus ponderosa forests of central Oregon. Forest Science 46:258-268.

Busso, C. A. and J. H. Richards. 1995. Drought and clipping effets on tiller demography and growth of two tussock grasses in Utah. Journal of Arid Environments 29:239-251.

Caudle, D., J. DiBenedetto, M. Karl, H. Sanchez, and C. Talbot. 2013. Interagency Ecological Site Handbook for Rangelands. Available at: http://jornada.nmsu.edu/sites/jornada.nmsu.edu/files/InteragencyEcolSiteHandbook.pdf. Accessed 4 October 2013.

Chambers, J., B. Bradley, C. Brown, C. D'Antonio, M. Germino, J. Grace, S. Hardegree, R. Miller, and D. Pyke. 2013. Resilience to stress and disturbance, and resistance to *Bromus tectorum* L. invasion in cold desert shrublands of western North America. Ecosystems 17:1-16.

Chambers, J. C., B. A. Roundy, R. R. Blank, S. E. Meyer, and A. Whittaker. 2007. What makes Great Basin sagebrush ecosystems invasible by *Bromus tectorum*? Ecological Monographs 77:117-145.

Clark, R. G., M. B. Carlton, and F. A. Sneva. 1982. Mortality of bitterbrush after burning and clipping in eastern Oregon. Journal of Range Management 35:711-714.

Clements, C. D. and J. A. Young. 2002. Restoring antelope bitterbrush. Rangelands 24:3-6.

Comstock, J. P. and J. R. Ehleringer. 1992. Plant adaptation in the Great Basin and Colorado Plateau. Western North American Naturalist 52:195-215.

Conrad, C. E. and C. E. Poulton. 1966. Effect of a wildfire on Idaho fescue and bluebunch wheatgrass. Journal of Range Management 19: 138-141.

Cook, J. G., T. J. Hershey, and L. L. Irwin. 1994. Vegetative response to burning on Wyoming mountain-shrub big game ranges. Journal of Range Management 47:296-302.

Currie, P. O., D. W. Reichert, J. C. Malechek, and O. C. Wallmo. 1977. Forage selection comparisons for mule deer and cattle under managed ponderosa pine. Journal of Range Management 30:352-356.

Daubenmire, R. 1970. Steppe vegetation of Washington. Technical bulletin. Washington Agriculture Experiment Station. 131 pp.

Daubenmire, R. 1975. Plant succession on abandoned fields, and fire influences in a steppe area in southeastern Washington. Northwest Science 49:36-48.

Dayton, W. 1937. Range Plant Handbook. USDA, Forest Service. Bull.

Dobrowolski, J. P., M. M. Caldwell, and J. H. Richards. 1990. Basin Hydrology and Plant Root Systems. In: C. B.

Osmand, L. F. Pitelka, G. M. Hildy [eds]. Plant biology of the basin and range. Ecological Studies. 80: 243-292

Eckert, R. E., Jr. and J. S. Spencer. 1987. Growth and reproduction of grasses heavily grazed under rest-rotation management. Journal of Range Management 40:156-159.

Evans, R. A. and J. A. Young. 1978. Effectiveness of rehabilitation practices following wildfire in a degraded big sagebrush-downy brome community. Journal of Range Management 31:185-188.

Everett, R. L. and K. Ward. 1984. Early plant succession on pinyon-juniper controlled burns. Northwest Science 58:57-68.

Fire Effects Information System [online] http://www.fs.fed.us/database/feis

Furniss, M. M. and W. F. Barr. 1975. Insects affecting important native shrubs of the northwestern United States. General Technical Report INT-19. Intermountain Forest and Range Experiment Station, U.S. Department of Agriculture, Forest Service. Ogden, UT. p. 68

Ganskopp, D. 1988. Defoliation of Thurber needlegrass: herbage and root responses. Journal of Range Management 41:472-476.

Garrison, G. A. 1953. Effects of clipping on some range shrubs. Journal of Range Management 6:309-317.

Hironaka, M., M. A. Fosberg, and A. H. Winward. 1983. Sagebrush-grass habitat types of southern Idaho. Bulletin Number 35. University of Idaho, Forest, Wildlife and Range Experiment Station, Moscow, ID.

Houghton, J.G., C.M. Sakamoto, and R.O. Gifford. 1975. Nevada's weather and climate. Special Publication 2. Nevada Bureau of Mines and Geology, Mackay School of Mines, University of Nevada, Reno, NV.

Jensen, M.E. 1990 Interpretation of environmental gradients which influence sagebrush community distribution in Northeastern Nevada. J. of Range Management 43:161-166.

Kerns, B. K., W. G. Thies, and C. G. Niwa. 2006. Season and severity of prescribed burn in ponderosa pine forests: implications for understory native and exotic plants. Ecoscience 13:44-55.

Kindschy, R. R., C. S. Undstrom, and J. D. Yoakum. 1982. Wildlife habitats in managed rangelands - the Great Basin of southeastern Oregon: pronghorns. Gen. Tech. Rep. PNW-GTR-145. Portland, OR. P. 18

Kitchen, S. G. and E. D. McArthur. 2007. Big and Black Sagebrush Landscapes. In: S. Hood, M. Miller [eds.]. Fire ecology and management of the major ecosystems of southern Utah. Gen. Tech. Rep. RMRMS-GTR-202. U.S. Department of Agriculture, Forest Service, Rocky Mountain Research Station, Fort Collins, CO. P. 73-95.

Koniak, S. 1985. Succession in pinyon-juniper woodlands following wildfire in the Great Basin. The Great Basin Naturalist 45:556-566.

Laycock, W. A. 1967. How heavy grazing and protection affect sagebrush-grass ranges. Journal of Range Management 20:206-213.

McConnell, B. R. and J. G. Smith. 1977. Influence of grazing on age-yield interactions in bitterbrush. Journal of Range Management 30:91-93.

Miller, R. F. and R. J. Tausch. 2000. The role of fire in pinyon and juniper woodlands: a descriptive analysis. In Proceedings of the Invasive Species Workshop: The Role of Fire in the Control and Spread of Invasive Species. Fire conference. P. 15-30

Mueggler, W. F. 1975. Rate and pattern of vigor recovery in Idaho fescue and bluebunch wheatgrass. Journal of Range Management 28:198-204.

Murray, R. 1983. Response of antelope bitterbrush to burning and spraying in southeastern Idaho. In: Tiedemann,

Arthur R.; Johnson, Kendall L., [eds.] Research and management of bitterbrush and cliffrose in western North America. General Technical Report INT-152. Ogden, UT: US Department of Agriculture, Forest Service, Intermountain Forest and Range Experiment Station p. 142-152.

National Oceanic and Atmospheric Administration. 2004. The North American Monsoon. Reports to the Nation. National Weather Service, Climate Prediction Center. Available online: http://www.weather.gov/

Richards, J. H. and M. M. Caldwell. 1987. Hydraulic lift: substantial nocturnal water transport between soil layers by Artemisia tridentata roots. Oecologia 73:486-489.

Robberecht, R. and G. Defossé. 1995. The relative sensitivity of two bunchgrass species to fire. International Journal of Wildland Fire 5:127-134.

Sheehy, D. P. and A. H. Winward. 1981. Relative palatability of seven Artemisia taxa to mule deer and sheep. Journal of Range Management 34:397-399.

Tausch, R. J. 1999. Historic pinyon and juniper woodland Ddvelopment. In: Proceedings: Ecology and management of pinyon–juniper communities within the interior west RMRS-P-9. Ogden, UT, USA: US Department of Agriculture, Forest Service, Rocky Mountain Research Station. P. 12-19.

Tausch, R. J. and N. E. West. 1988. Differential establishment of pinyon and juniper following fire. American Midland Naturalist 119:174-184.

Tisdale, E. W. and M. Hironaka. 1981. The sagebrush-grass region: A review of the ecological literature. University of Idaho, Forest, Wildlife and Range Experiment Station. Moscow, ID. P. 31

Uresk, D. W., J. F. Cline, and W. H. Rickard. 1976. Impact of wildfire on three perennial grasses in south-central Washington. Journal of Range Management 29:309-310.

Urness, P. J. 1965. Influence of range improvement practices on composition, production, and utilization of Artemisia Deer Winter Range in Central Oregon. Oregon State University.

USDA-NRCS Plants Database (Online; http://www.plants.usda.gov).

Vose, J. M. and A. S. White. 1991. Biomass response mechanisms of understory species the first year after prescribed burning in an Arizona ponderosa pine community. Forest Ecology and Management 40: 175-187.

Wood, M. K., Bruce A. Buchanan, & William Skeet. 1995. Shrub preference and utilization by big game on New Mexico reclaimed mine land. Journal of Range Management 48: 431-437.

Wright, H. A. and J. O. Klemmedson. 1965. Effect of fire on bunchgrasses of the sagebrush-grass region in southern Idaho. Ecology 46: 680-688.

Young, R. P. 1983. Fire as a vegetation management tool in rangelands of the intermountain region. In: S. Monsen, N. Shaw [eds.] Managing intermountain rangelands - Improvement of range and wildlife habitats. USDA, Forest Service. P. 18-31.

### Contributors

RK TK Stringham P Novak-Echenique

Approval Kendra Moseley, 4/25/2024

### Rangeland health reference sheet

Interpreting Indicators of Rangeland Health is a qualitative assessment protocol used to determine ecosystem condition based on benchmark characteristics described in the Reference Sheet. A suite of 17 (or more) indicators are typically considered in an assessment. The ecological site(s) representative of an assessment location must be known prior to applying the protocol and must be verified based on soils and climate. Current plant community cannot be used to identify the ecological site.

Author(s)/participant(s)	P NOVAK-ECHENIQUE
Contact for lead author	State Rangeland Management Specialist
Date	07/12/2012
Approved by	Kendra Moseley
Approval date	
Composition (Indicators 10 and 12) based on	Annual Production

### Indicators

- 1. **Number and extent of rills:** Rills are none to rare. A few rills can be expected on steeper slopes in areas subjected to summer convection storms or rapid spring snowmelt.
- 2. **Presence of water flow patterns:** Water flow patterns are none to rare but can be expected in areas recently subjected to summer convection storms or rapid snowmelt, usually on steeper slopes.
- 3. Number and height of erosional pedestals or terracettes: Some pedestalling of shallow rooted plants is expected.
- 4. Bare ground from Ecological Site Description or other studies (rock, litter, lichen, moss, plant canopy are not bare ground): Bare Ground 10-20% depending on the amount of surface cover of rock fragments.
- 5. Number of gullies and erosion associated with gullies: None
- 6. Extent of wind scoured, blowouts and/or depositional areas: None
- 7. Amount of litter movement (describe size and distance expected to travel): Fine litter (foliage from grasses and annual & perennial forbs) expected to move distance of slope length during intense summer convection storms or rapid snowmelt events. Persistent litter (large woody material) will remain in place except during large rainfall events.
- 8. Soil surface (top few mm) resistance to erosion (stability values are averages most sites will show a range of values): Soil stability values should be 3 to 6 on most soil textures found on this site. (To be field tested.)

- Soil surface structure and SOM content (include type of structure and A-horizon color and thickness): Surface structure is typically medium platy. Soil surface colors are pale browns and soils are typified by an ochric epipedon. Organic matter of the surface 2 to 4 inches is typically 1 to 2 percent dropping off quickly below. Organic matter content can be more or less depending on micro-topography.
- 10. Effect of community phase composition (relative proportion of different functional groups) and spatial distribution on infiltration and runoff: Perennial herbaceous plants (especially deep-rooted bunchgrasses [Idaho fescue & bluebunch wheatgrass] slow runoff and increase infiltration. Shrub canopy and associated litter break raindrop impact and provide opportunity for snow catch and accumulation on site.
- 11. Presence and thickness of compaction layer (usually none; describe soil profile features which may be mistaken for compaction on this site): Compacted layers are none. Subsoil argillic horizons are not to be interpreted as compacted.
- 12. Functional/Structural Groups (list in order of descending dominance by above-ground annual-production or live foliar cover using symbols: >>, >, = to indicate much greater than, greater than, and equal to):

Dominant: Deep-rooted, cool season, perennial bunchgrasses

Sub-dominant: low shrubs (low sagebrush)> associated shrubs >deep-rooted, cool season, perennial forbs>shallow-rooted, cool season, perennial bunchgrasses>fibrous, shallow-rooted, cool season, perennial and annual forbs.

Other:

Additional:

- 13. Amount of plant mortality and decadence (include which functional groups are expected to show mortality or decadence): Dead branches within individual shrubs common and standing dead shrub canopy material may be as much as 25% of total woody canopy; some of the mature bunchgrasses (<10%) have dead centers.</p>
- 14. Average percent litter cover (%) and depth ( in): Reference Plant Community: Under shrubs and between plant interspaces (<20%) and litter depth is < 1/4 inch.
- 15. Expected annual annual-production (this is TOTAL above-ground annual-production, not just forage annualproduction): For normal or average growing season (through mid-June) ± 300 lbs/ac; Spring moisture significantly affects total production
- 16. Potential invasive (including noxious) species (native and non-native). List species which BOTH characterize degraded states and have the potential to become a dominant or co-dominant species on the ecological site if their future establishment and growth is not actively controlled by management interventions. Species that become dominant for only one to several years (e.g., short-term response to drought or wildfire) are not invasive plants. Note that unlike other indicators, we are describing what is NOT expected in the reference state for the ecological site: Potential invaders include snakeweed, cheatgrass, and annual mustards.

17. **Perennial plant reproductive capability:** All functional groups should reproduce in average (or normal) and above average growing season years