

# Ecological site EX043B18I036 Droughty 19-24 inches precipitation zone Cryic Beaverhead Mountains

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#### General information

**Provisional**. A provisional ecological site description has undergone quality control and quality assurance review. It contains a working state and transition model and enough information to identify the ecological site.

#### **MLRA** notes

Major Land Resource Area (MLRA): 043B-Central Rocky Mountains

Major Land Resource Area (MLRA) 43B, Central Rocky Mountains, is an expansive area covering 76,000 square miles. This area is nearly equally divided amongst the states of Idaho (32 percent), Montana (38 percent), and Wyoming (30 percent) with a very small portion in extreme northeast Utah. The land area of this MLRA is largely federally owned and administered by the US Forest Service.

MLRA 43B is defined by its rugged mountains and contains both Montana's and Wyoming's tallest peaks. These mountains follow the Continental Divide and are the source for several major watersheds including the Columbia River, Snake River, Missouri River, and Yellowstone River. This water is primarily used in neighboring MLRAs as irrigation however it also supplies urban populations with domestic water and a growing recreational industry via fishing and boating.

The Central Rocky Mountains is used primarily for its forest/timber resources as well as used for grazing livestock. Also this area also provides for a vast recreational opportunities. The forests are dominated by coniferous trees such as ponderosa pine, Douglas fir, lodgepole pine, subalpine fir, and spruce. High elevation alpine forests may also include whitebark pine. The grazinglands of MLRA 43B are as variable and numerous as their forested counterparts with dominant grass and grass-like species including bluebunch wheatgrass, pinegrass, elk sedge, rough fescue, Idaho fescue, spike fescue, multiple bluegrass species, as well as numerous needlegrasses.

#### LRU notes

Land Resource Unit (LRU) 18 of MLRA 43B is also known as the Beaverhead Mountains LRU. It occupies approximately 4.2 million acres and generally resides on the border of Idaho and Montana. This LRU includes the Pioneer, Ruby, Beaverhead, Centennial, Gravelly, Snowcrest, Tobacco Root, Blacktail, and Tendoy Mountains. This LRU shares a boundary with MLRA 44B (Central Rocky Mountain Valleys), MLRA 12 (Lost River Valleys and Mountains), and MLRA 13 (Eastern Idaho Plateaus).

Like the majority of the LRUs within MLRA 43B, the geology of LRU 18 is best described as mixed; the geology is dominated by the Belt Supergroup. This geological supergroup is primarily sedimentary rocks with intrusive volcanics. The geological mix of this LRU includes Ravalli group; Piegan group; Missoula group; Tertiary volcanics; Mississippian group; Pennsylvanian group; Pre-belt gneiss, schist, and related rocks; Triassic and Permian; Montana group; Boulder batholith; Devonian Cambrian Undifferentiated; and alluvium. The geological formations of this LRU can contain significant amounts of mineral deposits such as gold, copper, silver, lead, platinum, palladium, and other precious metals.

Soils of this LRU are dominated by deep and very deep soils that formed on a variety of parent materials with shallow and mod-deep soils on ridges and near areas of rock outcrop. Moderate to very steep slopes dominate; however, broad, gently to moderately sloping areas also occur in places. Rock outcrop, rubble land, and surface rock fragments are common throughout these mountains with many soils expressing rock fragments within the

profile. Soil pH is highly variable particularly due to the mixed nature of the geology and varied precipitation amounts. Soils with lower precipitation, in this LRU, commonly have higher pH than those of higher precipitation. Most soils of this area fall into four soil orders: Entisols, Inceptisols, Mollisols, and Alfisols; soil profiles display very limited to highly developed horizonation and structure.

#### Classification relationships

EPA Ecoregions of Montana, Second Edition:

Level I: Northwestern Forested Mountains

Level II: Western Cordillera Level III: Middle Rockies

Level IV: Forested Beaverhead Mountains

Western Beaverhead Mountains

**Barren Mountains** 

Eastern Gravelly Mountains Pioneer-Anaconda Ranges

Eastern Pioneer Sedimentary Mountains

**Tobacco Root Mountains** 

Dry Gneissic-Shistose-Volcanic Hills

National Hierarchical Framework of Ecological Units (Forest Service):

Domain: Dry

Division: M330 – Temperate Steppe Division – Mountain Provinces

Province: M332 - Middle Rocky Mountain Steppe - Coniferous Forest - Alpine Meadow

Section: M332E – Beaverhead Mountains Section

National Vegetation Classification System (NVC):

2 Shrub & Herb Vegetation

2.B Temperate & Boreal Grassland & Shrubland

2.B.2 Temperate Grassland & Shrubland

2.B.2.Na Western North American Grassland & Shrubland

M048 Central Rocky Mountain Montane-Foothill Grassland & Shrubland

G267 Central Rocky Mountain Montane Grassland Group

G273 Central Rocky Mountain Lower Montane, Foothill & Valley Grassland Group

### **Ecological site concept**

Climatic Subset E:

- Frost Free Days (FFD) = 50 to 90 days
- Relative Effective Annual Precipitation (REAP) = 19 to 24 inches
- Moisture Regime: Typic, Ustic or Xeric (limited)
- Soil Temperature Regime: Cryic, warm

Climate Subset E typically resides at the mid elevations of MLRA 43B, LRU 18 near forested areas.

- · Site does not receive any additional water
- Soils are
- o Not saline or saline-sodic
- o Moderately deep, deep, or very deep
- o Typically less than 5 percent stone and boulder cover with a maximum of 15 percent
- o Skeletal (greater than 35 percent rock fragments by volume) at 10-20 inch control section
- o Not strongly or violently effervescent within surface mineral 4 inches; normally calcium carbonates increase with depth
- o Clay content is less than 32 percent in surface mineral 4 inches.
- o Soil does not have an argillic horizon with greater than 35 percent clay.
- Soil surface texture ranges from sandy loam to clay loam in surface mineral 4 inches
- Slope is less than 15 percent
- Parent material is primarily colluvium and slope alluvium.

#### **Associated sites**

EX043B18I038	Droughty Steep 19-24" PZ Cryic Beaverhead Mountains
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#### Similar sites

EX043B18I038	Droughty Steep 19-24" PZ Cryic Beaverhead Mountains
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Table 1. Dominant plant species

Tree	Not specified
Shrub	<ul><li>(1) Artemisia tridentata ssp. vaseyana</li><li>(2) Symphoricarpos albus</li></ul>
Herbaceous	<ul><li>(1) Achnatherum richardsonii</li><li>(2) Pseudoroegneria spicata</li></ul>

### Legacy ID

R043BZ436MT

### Physiographic features

This site exists on the middle third to top of gently sloping mountain slopes, ridges, fan remnants, hillslopes, and piedmont slope. Slopes vary from nearly level to less than 15 percent; however, the central concept of the site is between 4 and 10 percent slope.

Table 2. Representative physiographic features

Geomorphic position, mountains	<ul><li>(1) Center third of mountainflank</li><li>(2) Upper third of mountainflank</li><li>(3) Mountaintop</li></ul>
Landforms	<ul> <li>(1) Mountains &gt; Mountain slope</li> <li>(2) Mountains &gt; Ridge</li> <li>(3) Mountains &gt; Fan remnant</li> <li>(4) Hills &gt; Hillslope</li> <li>(5) Piedmont slope</li> </ul>
Elevation	1,707–2,652 m
Slope	0–15%
Aspect	W, NW, N, NE, E, SE, S, SW

### **Climatic features**

This ecological site exists in the Climate Subset E which is described as receiving 19 to 24 inches of Relative Effective Annual Precipitation (REAP) and having 50 to 90 Frost Free Days (FFD) though rarely exceeding 70 days. The average REAP for this Subset is 22 inches of precipitation with 65 FFD. This subset within the Beaverhead Mountains LRU has very limited climate station data due to the remote locations away from populated areas and Snowtel sites. Climate information was inferred from PRISM Climate Model Data.

Table 3. Representative climatic features

Frost-free period (characteristic range)	14-22 days
Freeze-free period (characteristic range)	36-39 days
Precipitation total (characteristic range)	457-508 mm
Frost-free period (actual range)	13-23 days
Freeze-free period (actual range)	35-40 days

Precipitation total (actual range)	457-508 mm
Frost-free period (average)	18 days
Freeze-free period (average)	38 days
Precipitation total (average)	483 mm

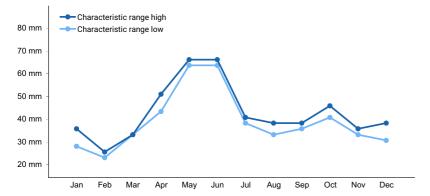


Figure 1. Monthly precipitation range

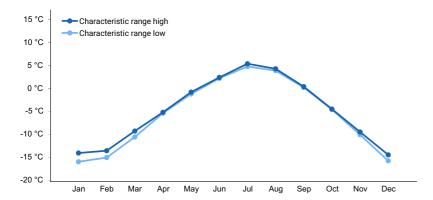


Figure 2. Monthly minimum temperature range

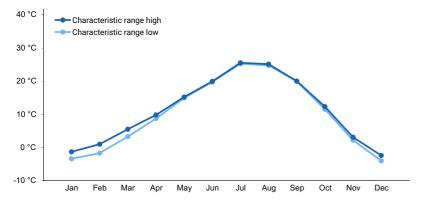


Figure 3. Monthly maximum temperature range

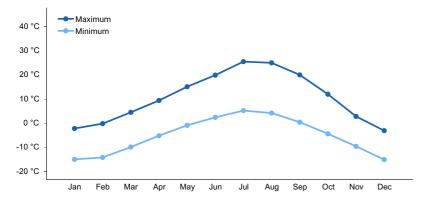


Figure 4. Monthly average minimum and maximum temperature

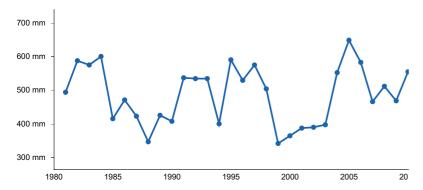


Figure 5. Annual precipitation pattern

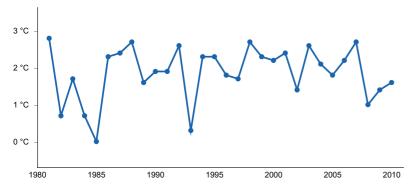


Figure 6. Annual average temperature pattern

#### **Climate stations used**

- (1) BIG SKY 2WNW [USC00240775], Gallatin Gateway, MT
- (2) LAKEVIEW [USC00244820], Lima, MT

#### Influencing water features

The Droughty ecological site is characterized by having greater than 35 percent rock fragments in the 10 to 20 inch depth of the soil profile. Often the rock fragment amounts increase with soil depth. These rock fragments reduce the overall soil water holding capacity. This is a water limited site with precipitation often infiltrating faster into the soil than running off. The exception is during high precipitation events associated with convective summer storms.

### Wetland description

Site is not associated with wetlands.

#### Soil features

The soils are moderately deep to very deep, moderately slow to moderately rapid permeability, and well drained.

These soils formed from colluvium and slope alluvium. The soil consists of loamy-skeletal material (averages greater than 35 percent rock fragments by volume in the 10-20 inch layer). This skeletal material decreases the water-holding capacity making the site easily susceptible to drought conditions. Typically soil surface textures consist of loam, gravelly loam, cobbly sandy loam, and cobbly loam, but may include clay loams. Common soil series in this ecological site are Sebud, Woodhurst, and Libeg. The soil color is typically dark; suggesting high organic matter content. Soil pH at the surface will be variable across the LRU and range from 6.2 to 7.8. Due to the mixed geology, pH may increase with depth. The Droughty ecological site will not be violently effervescent in the top 4 inches of soil. These soils may exist across multiple ecological sites due to natural variations in slope, texture, rock fragments, and pH. An onsite soil pit and most current Ecological Site Key are required to classify to an ecological site.



Figure 7. Droughty ecological site soil pit.

Table 4. Representative soil features

Parent material	<ul><li>(1) Colluvium–igneous, metamorphic and sedimentary rock</li><li>(2) Slope alluvium–igneous, metamorphic and sedimentary rock</li></ul>
Surface texture	(1) Cobbly, gravelly loam (2) Cobbly sandy loam
Family particle size	(1) Loamy-skeletal
Drainage class	Well drained
Permeability class	Moderately slow to rapid
Depth to restrictive layer	51–254 cm
Surface fragment cover <=3"	0–24%
Surface fragment cover >3"	0–15%
Available water capacity (0-101.6cm)	3.81–15.75 cm
Soil reaction (1:1 water) (0-10.2cm)	6–7.6

#### **Ecological dynamics**

The Droughty (Dr) ecological site is characterized by the production and composition of the Reference Plant Community, which is defined by soils, precipitation, and the temperature regime influencing the site. The Droughty ecological site is characterized by having skeletal soils greater than 20 inches deep and not being strongly or violently effervescent in the top 4 inches of the soil surface. This is an upland site that is heavily influenced by aspect and variations in plant communities do exist as a result. The Droughty Ecological Site extent in the Beaverhead Mountains LRU is limited and often resides on the edge of this MLRA.

In the Beaverhead Mountains LRU of MLRA 43B, Subset E is found where the Ustic soil moisture regime and a

Cryic temperature regime exist. In extremely limited portions of this site, a Xeric soil moisture regime may exist. This equates to a yearly representative value of 19 to 24 inches of relative effective annual precipitation and between 50 and 90 consecutive frost-free days though rarely exceeds 70. The majority of precipitation comes in April, May and June. Primary growth typically occurs between May and July with dominant plants that have adapted to these conditions. The warmer, drier section of this LRU primary growth typically occurs between April and June. A period of fall green-up can occur amongst this cool-season dominated plant community, if adequate precipitation is present.

The reference plant community is a midstatured, cool season bunchgrass community with bluebunch wheatgrass (*Pseudoroegneria spicata*), Richardson's needlegrass (*Achnatherum richardsonii*), Columbia needlegrass (Achnatherum neslonii), timber oatgrass (*Danthonia intermedia*), and Idaho fescue (*Festuca idahoensis*) as common dominant species. Subdominant species include mountain big sagebrush (*Artemisia tridentata* ssp vaseyana), snowberry (Symphoricarpos spp) with thickspike wheatgrass (*Elymus lanceolatus*), and green needlegrass (*Nassella viridula*). Though not always present in significant amounts, spike fescue (*Leucopoa kingii*) and rough fescue (*Festuca campestris*) are part of this ecological site.

The Soil and Vegetation Inventory of Near-pristine Sites of Montana suggests that the reference community is dominated primarily of bunchgrasses with less than 20 percent of the community being shrubs (Ross et al. 1973). Other historical accounts also suggest that sagebrush and shrubs were less extensive as part of the historic plant community for this area (Hayden 1873).

The Droughty ecological site occurs across a relatively large landscape with large variations within the plant community occurring due to geology, elevation, frost-free days, aspect, and relative effective annual precipitation. These variations tend to manipulate the amounts of individual species, however the core community is rarely changed.

Fire has historically been a powerful driver of this community. Some estimates of fire return may have been as long as 200 years in some sagebrush communities however a local study (Arno and Gruell 1982) suggests a fire return interval of 20 to 40 years. These frequent fires limited conifer expansion on this site, maintained potentially small aspen groves, and limited dense sagebrush communities.

This ecological site is largely intact and being primarily used for livestock grazing. Historical records indicate, prior to the introduction of livestock (cattle and sheep) during the late 1800's, elk and bison grazed this ecological site (Lesica and Cooper 1997). Due to the nomadic nature and herd structure of bison, areas that were grazed received periodic high intensity short duration grazing pressure. Meriwether Lewis documented that he was met by 60 Shoshone warriors on horseback in August 1805 and the Corps of Discovery was later supplied with horses by the same band of Shoshone (Moulton and Dunlay 1988). This suggests that the areas near the modern day towns of Twin Bridges, Bannack, Dillon, Grant, Lima, were grazed by an untold number of horses for years prior to the large introduction of cattle and sheep.

Livestock grazing has occurred on most of this ecological site for more than 150 years. The gold boom in the 1860s brought the first herds of livestock overland from Texas, and homesteaders began settling the area. Since the 1930's cattle production has dominated the livestock industry in the region (Wyckoff and Hansen 2001).

Dense clubmoss (*Selaginella densa*), in general, is a very minor component of Reference plant community of the Droughty Ecological Site. The conditions that created large cover classes of clubmoss on this site point to a history of continuous (yearlong) or moderate spring grazing use (Sturm 1954). While dense clubmoss provides soil stability on sites it exists, anecdotal accounts by some land managers suggest that it competes for the limited water resources into the upper soil profile which, in turn, restricts plant available water. However a study from Canada (Colberg and Romo 2003) indicates that the correlation between reduced plant available water and clubmoss cover is negligible. The correlation between reduced plant production may simply be competition for space though quantitative evidence is unavailable. Dense patches of clubmoss also inhibit seed contact with soil reducing seedling recruitment.

Some of the major invasive species that can occur on this site include (but not limited to) spotted knapweed (*Centaurea stoebe*), leafy spurge (*Euphorbia esula*), cheatgrass (*Bromus tectorum*), field brome (Bromus arevensis), yellow toadflax (*Linaria vulgaris*), Kentucky bluegrass (*Poa pratensis*), and dandelion (Taraxicum spp). Invasive weeds are beginning to have a high impact on the Droughty ecological site due to primarily human

impacts, mismanaged grazing, and urban development. Cheatgrass and spotted knapweed pose the largest threat to this ecological site.

#### Plant Communities and Transitional Pathways

A State and Transition Model for the Droughty ecological site is depicted below. Thorough descriptions of each state, transition, plant community, and pathway follow the model. This model is based on available experimental research, field data, field observations, and interpretations by experts. It is likely to change as knowledge increases.

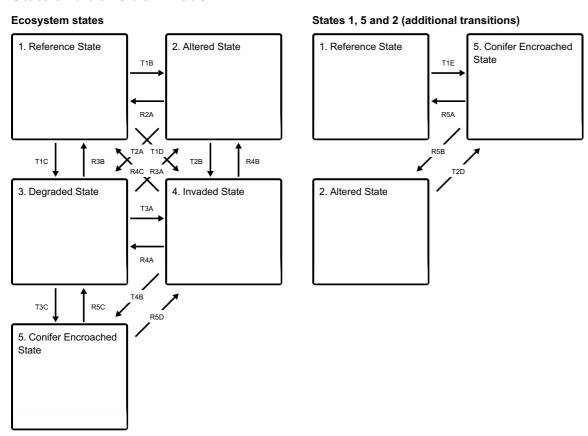
The plant communities within the same ecological site will differ across the MLRA due to the naturally occurring variability in weather, soils, and aspect. The biological processes on this site are complex; therefore, representative values are presented in a land management context. The species lists are representative and are not botanical descriptions of all species occurring, or potentially occurring, on this site. They are intended to cover the core species and known range of conditions and responses.

Both percent species composition by weight and percent canopy cover are referenced in this document. Most observers find it easier to visualize or estimate percent canopy for woody species (trees and shrubs). Canopy cover drives the transitions between communities and states because of the influence of shade, interception of rainfall and competition for available water. Species composition by dry weight remains an important descriptor of the herbaceous community and of the community as a whole. Woody species are included in species composition for the site. Calculating similarity index requires use of species composition by dry weight.

This State and Transition Model (STM) includes only rangeland communities and states. The converted communities are described in the Ecological Dynamics section (above).

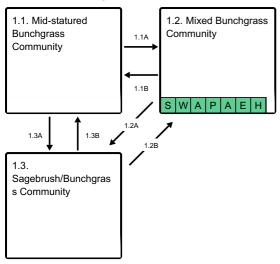
Although there is considerable qualitative experience supporting the pathways and transitions within the State and Transition Model (STM), no quantitative information exists that specifically identifies threshold parameters between grassland types and invaded types in this ecological site. For information on STMs, see the following citations: Bestelmeyer et al. 2003; Bestelmeyer and Brown 2005a; Bestelmeyer and Brown 2005b; Briske et al. 2008; Bestelmeyer et al. 2010; Bestelmeyer et al. 2016

#### State and transition model

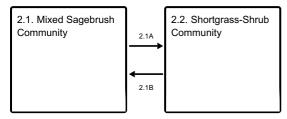


- **T1B** The trigger for this transition is improper grazing management and long-term drought leading to a decrease in bluebunch wheatgrass composition to 15 percent and reduction in total plant canopy cover.
- T1C The driver for this transition is improper grazing management, intense or repeated fires, or heavy human disturbance.
- T1D Recent dry climate cycles, repeated heavy grazing or intense human activities can open the interspaces of the bunchgrass community and allow for encroachment.
- T1E Conifer tree/shrub count exceeds two stems per acre. The trigger of crossing a threshold is the presence of seeds and/or other viable material of these tree species.
- R2A Recovery to the Reference State (1) will require reclamation efforts such as soil rebuilding, intensive mechanical and cultural treatments, and revegetation.
- T2A As improper grazing continues vigor of bunchgrasses will decrease, and the shorter grasses and shrubs will increase towards the Degraded State (3).
- T2B The trigger is the presence of seeds and other viable material of invasive species.
- **T2D** Conifer tree/shrub count exceeds two stems per acre. The trigger of crossing a threshold is the presence of seeds and other viable material of these tree species.
- **R3B** The drivers for the restoration pathway are removal of increaser species, restoration of native bunchgrass species, persistent management of invasives and shrubs, and proper grazing management.
- R3A If a sufficient amount of bunchgrass remains on the site, chemical application or biological control in conjunction with proper grazing management, can reduce the amount of shrubs and invasive species and restore the site to the Shortgrass Community (2.2).
- T3A The trigger is the presence of seeds or viable material of invasive species.
- T3C Conifer tree/shrub count exceeds two stems per acre. The trigger of crossing a threshold is the presence of seeds and other viable material of these tree species.
- **R4C** The drivers for the restoration pathway are removal of invasive species, restoration of native bunchgrass species, persistent management of invasive species, and proper grazing management.
- R4B The driver for the reclamation pathway is weed management with possible reseeding.
- R4A The driver for the reclamation pathway is weed management without reseeding.
- **T4B** Canopy cover of conifer tree/shrub cover exceeds two stems per acre. The threshold change is triggered by the presence of seeds and other viable material of invasive species.
- **R5A** Recovery requires a more intense mechanical removal of trees/shrubs. Reseeding and increased post treatment grazing management may be necessary.
- **R5B** Recovery requires a more intense mechanical removal of trees/shrubs. Reseeding and increased post treatment grazing management may be necessary.
- **R5C** Recovery requires a more intense mechanical removal of trees/shrubs. Reseeding and increased post treatment grazing management may be necessary.
- **R5D** Recovery requires a more intense mechanical removal of trees/shrubs. Reseeding and increased post treatment grazing management may be necessary.

### State 1 submodel, plant communities

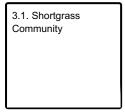


### State 2 submodel, plant communities



2.1A - The driver for community shift 2.1A is continued overgrazing.

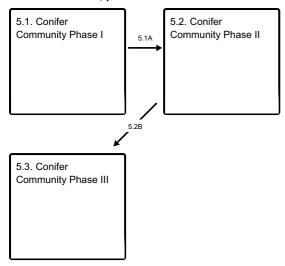
#### State 3 submodel, plant communities



#### State 4 submodel, plant communities



#### State 5 submodel, plant communities



- **5.1A** The driver for this pathway is primarily lack of fire; however, heavy grazing (utilization that exceeds 50 percent) can help reduce herbaceous competition and expose soil for seed contact.
- **5.2B** The driver for this pathway is primarily lack of fire; however, heavy grazing (can help reduce herbaceous competition and expose soil for seed contact.

## State 1 Reference State

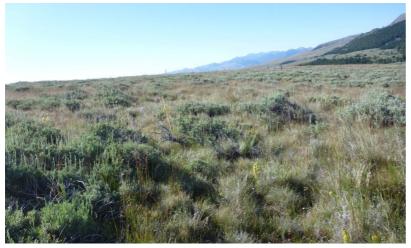


Figure 8. Picture of Reference State in a transition from Community 1.1 to 1.2.

The Reference State of this ecological site consists of 3 potential plant communities 1.1 Mid-statured Bunchgrass Community, 1.2 Mixed Bunchgrass Community, and 1.3 Sagebrush/Bunchgrass Community. These are described further in their respective community details but are generally characterized by mid-statured, cool-season bunchgrass communities with moderate shrub production. Due to the increased spring and summer precipitation, these communities often resistant to disturbance. Community 1.1 is dominated by mid-statured, cool-season bunchgrasses such as bluebunch wheatgrass, Richardon's needlegrass, and Columbia needlegrass with a mixed shrub component. Community 1.1 is considered the Reference Plant Community while Community 1.2 sees a reduction in mid-statured bunchgrasses and an increase in short bunchgrasses and forbs. Community 1.3 is codominated by bluebunch wheatgrass and Idaho fescue with mountain big sagebrush as the dominant shrub.

## Community 1.1 Mid-statured Bunchgrass Community

This community occurs in areas with low grazing use. Prior to the introduction of expansive livestock herds, basin wildrye may have dominated this ecological site, as suggested by a Forest Service Technical Report (Gruell 1983), however quantitative evidence is lacking and relict sites are very few and small in size. In the Mid-statured Bunchgrass Community, bluebunch wheatgrass (*Pseudoroegneria spicata*), slender wheatgrass (*Elymus* 

trachycaulus), Idaho fescue (Festuca idahoensis), Richardson's needlegrass (Achnatherum richardsonii), and Columbia needlegrass (Acnatherum nelsonii) are dominant. Basin wildrye (Elymus cinerus), green needlegrass (Nasella viridula), and Cusick's bluegrass (Poa cusickii) will be subordinate grasses. Shrub species such as common snowberry (Symphoricarpus albus), mountain big sagebrush (Artemisia tridentata ssp vaseyana), and spineless horsebrush (Tetradymia canescence) remain a minor part of the community (Hayden 1873). Sandberg bluegrass (Poa secunda) and dryland sedges will also be present. Midstatured bungrasses lack resistance to grazing during the critical growing season and will decline in vigor and production if grazed in that time frame more than one year in three (Wilson et al. 1960). The critical growing period for these mid-statured bunchgrasses is late spring (June). This community is moderately resilient and will return to dynamic equilibrium following a relatively short period of stress (such as drought or short-term improper grazing), provided a return of favorable or normal growing conditions, and properly managed grazing. Improper grazing is defined as grazing utilization that exceeds moderate use (less than 50 percent grazing use) or multiple grazing events of a plant in the same growing season without rest. As discussed in the Ecological Dynamics section, natural fire regime restricted shrubs to relatively small portions of Community 1.1. Mountain big sagebrush remains the dominant shrub with canopy cover less than 10 percent. Shrub species not listed above may also include, but not limited to chokecherry (*Prunus virginiana*), shrubby cinquefoil (Dasiphora fruiticosa), silver sagebrush (Artemisia cana), winterfat (Krascheninnikovia lanata), tarragon (Artemisia drucunculus), and fringed sagewort (Artemisia frigida).

#### **Dominant plant species**

- mountain big sagebrush (Artemisia tridentata ssp. vaseyana), shrub
- common snowberry (Symphoricarpos albus), shrub
- spineless horsebrush (*Tetradymia canescens*), shrub
- yellow rabbitbrush (Chrysothamnus viscidiflorus), shrub
- bluebunch wheatgrass (Pseudoroegneria spicata), grass
- Idaho fescue (Festuca idahoensis), grass
- Richardson's needlegrass (Achnatherum richardsonii), grass
- lupine (Lupinus), other herbaceous
- common yarrow (Achillea millefolium), other herbaceous
- beardtongue (*Penstemon*), other herbaceous
- western stoneseed (*Lithospermum ruderale*), other herbaceous
- elkweed (*Frasera speciosa*), other herbaceous

Table 5. Annual production by plant type

Plant Type	Low (Kg/Hectare)	Representative Value (Kg/Hectare)	High (Kg/Hectare)
Grass/Grasslike	1054	1412	2074
Forb	224	303	448
Shrub/Vine	224	291	359
Tree	-	11	56
Total	1502	2017	2937

#### Table 6. Ground cover

Tree foliar cover	0-1%
Shrub/vine/liana foliar cover	0-15%
Grass/grasslike foliar cover	35-50%
Forb foliar cover	5-15%
Non-vascular plants	0%
Biological crusts	0-3%
Litter	25-35%
Surface fragments >0.25" and <=3"	0-10%
Surface fragments >3"	0-5%

Bedrock	0%
Water	0%
Bare ground	0%

Table 7. Soil surface cover

Tree basal cover	0%
Shrub/vine/liana basal cover	1-5%
Grass/grasslike basal cover	5-10%
Forb basal cover	1-3%
Non-vascular plants	0%
Biological crusts	0-3%
Litter	25-35%
Surface fragments >0.25" and <=3"	0-10%
Surface fragments >3"	0-5%
Bedrock	0%
Water	0%
Bare ground	0-10%

## Community 1.2 Mixed Bunchgrass Community



Idaho fescue and bluebunch wheatgrass share dominance in the Mixed Bunchgrass Community (1.2) with a decline of slender wheatgrass and decreaser needlegrasses. Other grass species, which are more tolerant to grazing and are likely to increase compared to the Midstatured Bunchgrass Community, including Sandberg bluegrass (Poa secunda), Cusick's bluebrass (Poa cusickii), prairie Junegrass (Koeleria macrathana), western/thickspike wheatgrass (Pascopyrum smithii, Elymus lanceolatus), and needle and thread (Hesperostipa comata). Some increaser forbs species include western yarrow (Achillea millefolium), Hoods phlox (Phlox hoodii), lupine (Lupinus spp), mountain death camas (Zigadenus elegans), scarlet globemallow (Sphaeralcea coccinea), hairy goldenaster (Heterotheca villosa), prairie smoke (Geum triflorum), and pussytoes (Antennaria spp.). Fringed sagewort may also increase under prolonged drought or heavy grazing and can respond to precipitation that falls in July and August. Idaho fescue tolerates grazing pressure better than bluebunch wheatgrass, Richardson's needlegrass, and Columbia needlegrass. The growing point for bluebunch wheatgrass is several inches above the ground, making them very susceptible to continued close grazing (Smoliack, et al 2006), while Idaho fescue's growing point tend to be near the plant base. These grasses increases in species composition when more palatable and less grazing tolerant plants decrease due to overgrazing, drought, or intense fire. Heavy, continuous grazing will reduce plant cover, litter, and soil organic matter. Timing of grazing is important on this site as grazing decreaser bunchgrasses in summer season has proven to negatively impact seasonal regeneration, especially on the drier sites. Bare ground will increase which may increase soil erosion. Litter and mulch will be reduced as plant cover declines. As long as bluebunch wheatgrass is still a dominant species of total biomass production, the site can return to the Midstatured Bunchgrass Community (Pathway 1.2A) under proper grazing management and favorable growing conditions. In this community, proper grazing management is a grazing rotation that favors decreaser bunchgrasses. This grazing management plan would include conservative grazing use with light to moderate grazing use during the critical growing season which is early to mid summer (boot stage of the plant) or preferably dormant grazing use. (Wilson et al. 1966; McLean and Wikeem 1985) Over time, Idaho fescue, native bluegrasses, and western wheatgrass will increase until they make up the majority of species composition. Once midstatured grasses have been reduced to less than 40 percent by dry weight and mountain big sagebrush canopy cover exceeds 15 percent, it may be difficult for the site to recover to Midstatured Bunchgrass Community (1.1). The risk of soil erosion increases when canopy cover decreases. As soil conditions degrade, there will be loss of organic matter, reduced litter, and reduced soil fertility. Degraded soil conditions increase the difficulty of reestablishing bluebunch wheatgrass preventing the return to the Midstatured Bunchgrass Community (1.1). If unmanaged grazing continues, increaser species such as needle and thread Sandberg bluegrass, and native forb species will become more dominate and this triggers the change to the Altered State (2) or the Degraded State (3). Unless the Mixed Bunchgrass Community (1.2) crosses the threshold into the Mixed Sagebrush Community (2.1) or the Invaded Community (4.1), this community can be managed toward the Midstatured Bunchgrass Community (1.1) using prescribed grazing and strategic weed control. It may take several years to achieve this recovery, depending on growing conditions, vigor of bluebunch and other midstatured bunchgrass plants, and the aggressiveness of the weed treatments (if necessary).

#### **Dominant resource concerns**

- Sheet and rill erosion
- Plant productivity and health
- Plant structure and composition
- Terrestrial habitat for wildlife and invertebrates

## Community 1.3 Sagebrush/Bunchgrass Community



The Sagebrush/Bunchgrass Community is best described as a community with a high mountain big sagebrush canopy cover that limits an understory of the bunchgrass community. This community is often a result of long term seasonal use with a lack of fire. Of the communities in the Reference State the Sagebrush/Bunchgrass Community is at highest risk for conifer encroachment. Over time, Idaho fescue, native bluegrasses, and thickspike wheatgrass will increase until they make up the majority of species composition. Once midstatured grasses have been reduced to less than 20 percent by dry weight and mountain big sagebrush canopy cover exceeds 20 percent, it may be difficult for the site to recover to Midstatured Bunchgrass Community (1.1). The risk of soil erosion increases when canopy cover decreases. As soil conditions degrade, there will be loss of organic matter, reduced litter, and reduced soil fertility. Degraded soil conditions increase the difficulty of reestablishing bluebunch wheatgrass preventing the return to the Midstatured Bunchgrass Community (1.1). The Sagebrush/Bunchgrass Community (1.3) is considered an At-Risk Plant Community for this ecological site. When overgrazing continues increaser species such as needle and thread and native forb species will become more dominate and this triggers the change to the Altered State (2) or the Degraded State (3). Until the Sagebrush/Bunchgrass Community (1.3) crosses the threshold into the Mixed Sagebrush Community (2.1) or the Invaded Community (4.1), this community can be managed toward the

Midstatured Bunchgrass Community (1.1) using prescribed grazing, prescribed fire, and strategic weed control. It may take several years to achieve this recovery, depending on growing conditions, vigor of bluebunch and other midstatured bunchgrass plants, and the aggressiveness of the weed treatments (if necessary).

### Pathway 1.1A Community 1.1 to 1.2

Midstatured bunchgrasses lose vigor with improper grazing (see definition below) or extended drought. When vigor declines in response to disturbances such as these, bunchgrass basal areas will decrease in size while species with higher grazing tolerance (primarily Idaho fescue on this site) increase in vigor and production as they access the resources previously used by decreaser bunchgrasses. The decrease of midstatured bunchgrasses to less than 40 percent composition and an increase of mountain big sagebrush canopy cover to 15 percent indicates that the plant community has shifted to the Mixed Bunchgrass Community (1.2). The driver for community shift 1.1A is improper grazing management or prolonged drought. This shift is triggered by the loss of vigor of native decreaser needlegrasses and increased bare ground associated with prolonged drought coupled with improper grazing. Blaisdell (1958) stated that drought and warmer than normal temperatures are known to advance plant phenology by as much as one month. During drought years, plants may be especially sensitive or in a critical stage of development earlier than expected. Improper grazing on this ecological site is defined as grazing utilization that exceeds moderate use (less than 50 percent grazing use) and/or multiple grazing events of a plant in the same growing season without rest. Associated with grazing intensity is timing of grazing. The critical grazing period for this site is late spring (approximately boot stage on grasses).

### Pathway 1.3A Community 1.1 to 1.3

Midstatured bunchgrasses lose vigor with improper grazing (see definition below), extended drought, and/or long term suppression of fire. When vigor declines in response to disturbances such as these, bunchgrass basal areas will decrease in size while species with higher grazing tolerance (primarily Idaho fescue on this site) increase in vigor and production as they access the resources previously used by decreaser bunchgrasses. The decrease of midstatured bunchgrasses and increase of mountain big sagebrush canopy cover over 20 percent indicates that the plant community has shifted to the Sagebrush/Bunchgrass Community (1.3). The driver for community shift 1.1A is improper grazing management or long term suppression of fire. This shift is triggered by the loss of vigor of native decreaser needlegrasses and increased bare ground associated with drought and improper grazing. Blaisdell (1958) stated that drought and warmer than normal temperatures are known to advance plant phenology by as much as one month. During drought years, plants may be especially sensitive or in a critical stage of development earlier than expected. Improper grazing on this ecological site is defined as grazing utilization that exceeds moderate utilization (approximately 50 percent grazing use) without rest. Timing of grazing events is also important to balancing resource conditions. The critical grazing period for this site is late spring (approximately boot stage on decreaser, cool season grasses). Grazing methodologies can vary greatly and no one management plan will work best so each strategy should be evaluated to match the current science based information available. Suppression of fire is often the primary trigger on this ecological site as a result of successful public campaign that began in the early 1900s. The increased fire regime has resulted in dense, single aged, canopy cover of sagebrush. This often associated with in an increase of surficial bare ground where coniferous tree seedlings can establish. The Sagebrush/Bunchgrass Community (1.3) will often transition to the Conifer Encroached State (5) with continued suppression of fire.

### Pathway 1.1B Community 1.2 to 1.1

The Mixed Bunchgrass Community (1.2) will return to the Midstatured Bunchgrass Community (1.1) with proper grazing management and appropriate grazing intensity. Favorable moisture conditions will facilitate or accelerate this transition. It may take several years of favorable conditions for the community to transition back to a decreaser bunchgrass dominated state. The driver for this community shift (1.1B) is increased vigor of midstatured bunchgrasses to the point that it represents more than 40 percent of species composition. The trigger for this shift is the change in grazing management favoring decreaser bunchgrasses. In general, conservative grazing management styles such as deferred or rest rotations utilizing moderate grazing (less than 50 percent grazing use) coupled with favorable growing conditions like cool, wet springs are these triggers. These systems tend to promote

increases in soil organic matter which, in turn, also promotes microfauna and an increase infiltration rates. Inversely, long periods of rest at a time when this state is considered to be stable may not result in an increase in bluebunch wheatgrass and similar decreaser bunchgrasses. It has been suggested (Noy-Meir 1975) that these long periods of rest or underutilization may actually drive the system to a lower level of stability by creating large amounts of standing biomass, dead plant caudex centers, and gaps in the plant canopy.

#### **Conservation practices**

**Prescribed Burning** 

**Prescribed Grazing** 

### Pathway 1.2A Community 1.2 to 1.3



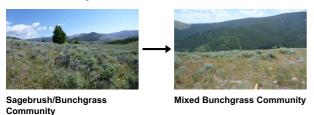
Midstatured bunchgrasses lose vigor with improper grazing (see definition below), extended drought, and/or long term suppression of fire. When vigor declines in response to disturbances such as these, bunchgrass basal areas will decrease in size while species with higher grazing tolerance (primarily Idaho fescue on this site) increase in vigor and production as they access the resources previously used by decreaser bunchgrasses. The decrease of midstatured bunchgrasses and increase of mountain big sagebrush canopy cover over 20 percent indicates that the plant community has shifted to the Sagebrush/Bunchgrass Community (1.3). The driver for community shift 1.2A is improper grazing management or long term suppression of fire. This shift is triggered by the loss of vigor of native decreaser needlegrasses and increased bare ground associated with drought and improper grazing. Blaisdell (1958) stated that drought and warmer than normal temperatures are known to advance plant phenology by as much as one month. During drought years, plants may be especially sensitive or in a critical stage of development earlier than expected. Improper grazing on this ecological site is defined as grazing utilization that exceeds moderate utilization (approximately 50 percent grazing use) without rest. Timing of grazing events is also important to balancing resource conditions. The critical grazing period for this site is late spring (approximately boot stage on decreaser, cool season grasses). Grazing methodologies can vary greatly and no one management plan will work best so each strategy should be evaluated to match the current science based information available. Suppression of fire is often the primary trigger on this ecological site as a result of successful public campaign that began in the early 1900s. The increased fire regime has resulted in dense, single aged, canopy cover of sagebrush. This often associated with in an increase of surficial bare ground where coniferous tree seedlings can establish. The Sagebrush/Bunchgrass Community (1.3) will often transition to the Conifer Encroached State (5) with continued suppression of fire.

### Pathway 1.3B Community 1.3 to 1.1

The Sagebrush/Bunchgrass Community (1.3) will return to the Mistatured Bunchgrass Community (1.1) with proper grazing management with appropriate grazing intensity and reintroduction of the natural fire regime. Favorable moisture conditions will facilitate or accelerate this transition. It may take several years of favorable conditions for the community to transition back to a decreaser bunchgrass dominated state. The driver for this community shift (1.3B) is increased vigor of midstatured bunchgrasses to the point that it represents more than 30 percent of species composition. The trigger for this shift is the change in grazing management favoring mistatured bunchgrasses. In general, conservative grazing management styles such as deferred or rest rotations utilizing moderate grazing (less than 50 percent grazing use) coupled with favorable growing conditions like cool, wet springs are these triggers. These systems tend to promote increases in soil organic matter which, in turn, also promotes microfauna and an increase infiltration rates. Inversely, long periods of rest at a time when this state is considered to be stable may not result in an increase in bluebunch wheatgrass and similar midstatured bunchgrasses. It has been suggested (Noy-Meir 1975) that these long periods of rest or underutilization may

actually drive the system to a lower level of stability by creating large amounts of standing biomass, dead plant caudex centers, and gaps in the plant canopy.

### Pathway 1.2B Community 1.3 to 1.2



The Sagebrush/Bunchgrass Community (1.3) will return to the Mixed Bunchgrass Community (1.2) with proper grazing management with appropriate grazing intensity and reintroduction of the natural fire regime. Favorable moisture conditions will facilitate or accelerate this transition. It may take several years of favorable conditions for the community to transition back to a decreaser bunchgrass dominated state. The driver for this community shift (1.2B) is increased vigor of midstatured bunchgrasses to the point that it represents more than 30 percent of species composition. The trigger for this shift is the change in grazing management favoring mistatured bunchgrasses. In general, conservative grazing management styles such as deferred or rest rotations utilizing moderate grazing (less than 50 percent grazing use) coupled with favorable growing conditions like cool, wet springs are these triggers. These systems tend to promote increases in soil organic matter which, in turn, also promotes microfauna and an increase infiltration rates. Inversely, long periods of rest at a time when this state is considered to be stable may not result in an increase in bluebunch wheatgrass and similar midstatured bunchgrasses. It has been suggested (Noy-Meir 1975) that these long periods of rest or underutilization may actually drive the system to a lower level of stability by creating large amounts of standing biomass, dead plant caudex centers, and gaps in the plant canopy.

### State 2 Altered State



This state is characterized by having less than 15 percent bluebunch wheatgrass by dry weight. It is represented by two communities that differ in the percent composition of Idaho fescue and needle and thread, and overall production. Production in this state can be similar to the Reference State (1). Some native plants tend to increase under prolonged drought and heavy grazing practices. A few of these species may include Idaho fescue, Sandberg bluegrass, Cusick's bluegrass, scarlet globemallow, hairy goldenaster, and fringed sagewort. Poisonous plants such as lupine (Lupinus spp) and mountain deathcamas (*Zigadenus elegans*) may also increase under these conditions.

## Community 2.1 Mixed Sagebrush Community



Idaho fescue is dominant in the Mixed Sagebrush Community (2.1). Mid-statured bunchgrasses comprise up to 15 percent of species composition by dry weight. The remaining mid-statured bunchgrasses tend to be scattered and low in vigor. Increaser and invader species will be more common. Increaser forb species include hairy goldenaster, death camas, goldenrod, stonecrop, lupine, and yarrow. It is not uncommon for a minor component of invader species such as dandelion and goatsbeard to be present. This creates more competition for the mid-statured bunchgrasses; making it difficult for them to quickly respond to a change in grazing management alone. Therefore, an input of energy is required for the community to return to the Reference State (1). Wind and water may be eroding soil from the plant interspaces. Soil fertility is reduced and soil surface erosion resistance has declined compared to the Reference State (1). Long-term grazing mismanagement with continuous growing-season pressure will reduce total productivity of the site and lead to an increase of bare ground. Suppression of fire can also promote shrub growth, increasing plant interspaces. Once plant cover is reduced, the site is more susceptible to erosion and degradation of soil properties. Soil erosion or reduced soil fertility will result in reduced plant production. This soil erosion or loss of soil fertility indicates the transition to the Altered State (2), because it has crossed a threshold requiring input of energy to return to the Reference State (1). Transition to the Mixed Sagebrush Community Community (2.1) may be exacerbated by extended drought conditions. Big sagebrush steppe communities historically had moderate fuel loadings and were characterized by 10 to 70 year fire return intervals (often less than 30 years), that produced a mosaic of burned and unburned lands (Bunting, et.al. 1987). Following fire on the fine-textured soils, the perennial bunchgrasses recovered in a few years and were present to fuel a subsequent fire. Conversely, extensive wildfires burning under hot-dry conditions would have resulted in nearly complete destruction of scattered sagebrush (Arno and Gruell 1983). Winterfat is tolerant of low intensity fire but can be removed by a hot fire (Pellant 1984). This community crossed a threshold from the Reference State (1) due to the erosion of soil, altered vegetative composition, loss of soil fertility, or degradation of soil conditions. This results in a critical shift in the ecology of the site. The effects of soil erosion can alter the hydrology, soil chemistry, soil microorganisms, and soil structure to the point where intensive restoration is required to restore the site to another state or community. Changing grazing management, alone, may not create sufficient improvement to restore the site if decreaser species are greatly reduced. Dormaar (1997) stated that with decreased grazing pressure, a needle and thread plant community did not change species composition but the content of the soil carbon increased. It will require a considerable input of energy to move the site back to the Reference State (1). This state has lost soil or vegetation attributes to the point that recovery to the Reference State (1) will require reclamation efforts, i.e., soil rebuilding, intensive mechanical treatments, and reseeding. The transition to this state could result from overgrazing and fire suppression, especially repeated early season grazing coupled with extensive drought. If heavy grazing continues, plant cover, litter, and mulch will continue to decrease and bare ground will increase exposing the soil to accelerated erosion. Litter and mulch will move off-site as plant cover declines. The Mixed Sagebrush Community will then shift to a Shortgrass-Shrub Community (2.2). Continued overgrazing will drive the community to a Degraded State (3). Introduction or expansion of invasive species will further drive the plant community to the Invaded State (4).

Community 2.2 Shortgrass-Shrub Community



With continued mismanagement of grazing, especially coupled with prolonged drought, Idaho fescue will decrease in vigor. The bunchgrasses will decline in production as plants die or become smaller, and species with higher grazing tolerance (such as western wheatgrass) increase in vigor and production as they respond to resources previously used by the bunchgrasses. These less desirable, shorter rooted species will become codominant with the bunchgrasses. Mid-statured grasses such as bluebunch wheatgrass, Columbia needlegrass, and green needlegrass will exist within taller shrubs, where they are protected from grazing animals, and only a few plants exist outside of the shrub canopy. Shrubs will become more competitive for limited moisture as bare ground and soil erosion increase. This state may exhibit conditions where livestock and wildlife are overconsuming shrubs.

### Pathway 2.1A Community 2.1 to 2.2



**Mixed Sagebrush Community** 

**Shortgrass-Shrub Community** 

The driver for community shift 2.1A is continued overgrazing. Overgrazing on this site is defined by grazing utilization over 60 percent or season long grazing where livestock can graze regrowth. This shift is triggered by continued loss of bunchgrass vigor, especially the remaining Idaho fescue and needlegrasses. The mid- and short-statured grasses will become more competitive and become co-dominant with the bunchgrasses. Shrubs will increase to 25 to 30 percent canopy cover. A shift in shrub dominance from big sagebrush to three-tip sagebrush (*Artemisia tripartita*) may also occur.

### Pathway 2.1B Community 2.2 to 2.1



Shortgrass-Shrub Community Mixed Sagebrush Community

If grazing management is implemented, Idaho fescue may regain its vigor and move towards the Mixed Sagebrush Community (2.1). This will give grasses an advantage over invading shrubs before too much competition takes place. Since the transition from Plant Community 2.1 to Plant Community 2.2 was likely caused by repeated heavy utilization, a conservative grazing plan, utilization is reduced at or below moderate levels with rest or deferment incorporated, will aid in this transition. Van Poolen and Lacey (1979) found that forage production increased by an average of 35 percent on western ranges when converting from heavy utilization to moderate utilization (less than 50 percent). Shrub removal and favorable growing conditions can accelerate this process. If the site contains excess mountain big sagebrush, a low intensity fire or mechanical treatment (Wambolt 1986) could reduce shrub competition and allow for increased vigor and the reestablishment of grass species.

## State 3 Degraded State

Degraded State is characterized by a near complete lack of mid-statured bunchgrasses. Idaho fescue, Sandberg bluegrass and prairie Junegrass are dominant grasses. Increaser shrubs nearly replace larger shrub species. Remaining larger shrub species are heavily hedged. This state is likely a terminal state (e.g. restoration is impossible or unsuccessful without major financial and energy inputs).

## Community 3.1 Shortgrass Community

Community Phase 3.1 is characterized by soil surface degradation and little plant soil surface cover. Soil loss continues and subsequent loss of soil organic matter creates conditions where native perennial grasses are reduced. Grass and forb cover may be sparse with obvious rill erosion between plant bases. Extensive dense clubmoss patches may also exist. Needle and thread and Idaho fescue exist in small patches. This could occur due to overgrazing (failure to adjust stocking rate to declining forage production due to increased invasive dominance), long-term lack of fire (if dense Wyoming big sagebrush occurs), or introduction of invasive species. In the most severe stages of degradation, there is a significant amount of bare ground, and large gaps occur between plants. Potential exists for soils to erode to the point that irreversible damage may occur. This is a critical shift in the ecology of the site. Soil erosion combined with lack of organic matter deposition due to sparse vegetation create changes to the hydrology, soil chemistry, soil microorganisms, and soil structure to the point where intensive restoration is required to restore the site to another state or community. Changing management (i.e., improving grazing management) cannot create sufficient change to restore the site. The forb component changes to be dominated by spiny phlox (*Phlox hoodii*) and shrub canopy cover is usually greater than 20 percent. Big sagebrush is replaced with a dominant community of low sagebrush (Artemisia arbuscula), broom snakeweed, rubber rabbitbrush, fringed sagewort, and plains pricklypear cactus. This community phase has lost soil or vegetation attributes to the point that recovery to the Reference State will require extensive reclamation efforts, i.e. soil rebuilding, intensive mechanical treatments, and reseeding. Grazing management such as a rest rotation will not return this state to Reference without reclamation efforts. This plant community may be in a terminal state that will not return to the Reference State because of degraded soil conditions and loss of higher successional native plant species. Key factors of the approach of a transition are decrease in grass canopy cover and production, increase of shrub canopy cover, increases in mean bare patch size, increases in soil crusting, decreases in cover of cryptobiotic crusts, decreases in soil aggregate stability, or evidence of erosion including water flow patterns and litter movement.

### State 4 Invaded State

The Invaded State is identified by being in the exponential growth phase of invader abundance where control is a priority. Dominance (or relative dominance) of noxious/invasive species reduces species diversity, forage production, wildlife habitat, and site protection. A level of 10 percent by weight invasive species composition by dry weight indicates the threshold of when substantial energy input, such as herbicide or mechanical removal, will be required to create a shift to the grassland state; even with a return to proper grazing management or favorable growing conditions. Prescriptive grazing that specifically focuses on grazing of the invasive plant can be used to manage invasive species. In some instances, carefully targeted grazing (sometimes in combination with other treatments) can reduce or maintain species composition of invasive species.

## Community 4.1 Invaded Community

Communities in this state may be structurally indistinguishable from the Reference State except that invasive/noxious species exceed 10 percent of species composition by dry weight. This state may also include a community similar to the Degraded State (3) except that invasive/noxious species exceed 10 percent of species composition by dry weight. Although there is no research to document the level of 10 percent, this is estimated to be the point in the invasion process following the lag phase based on interpretation of Masters and Sheley, 2001. For aggressive invasive species (i.e., spotted knapweed) the threshold could be less than 10 percent. Early in the

invasion process there is a lag phase where the invasive plant populations remain small and localized for long periods before expanding exponentially (Hobbs and Humphries 1995). Production in the invaded community may vary greatly based on the severity of invasion and potential subsequent effects on species composition. A site dominated by Kentucky bluegrass or spotted knapweed, where soil fertility and chemistry remain near reference, may have production near that of the reference community. A site with degraded soils and an infestation of cheatgrass may produce only 10 to 20 percent of the reference community. Once invasive species dominate the site, either in species composition by weight or in their impact on the community the threshold has been crossed to the Invaded State (4). As invasive species such as spotted knapweed, cheatgrass, and leafy spurge become established, they become very difficult to eradicate. Therefore considerable effort should be placed in preventing plant communities from crossing the threshold to the Invaded State (4) through early detection and proper management. Preventing new invasions is by far the most cost-effective control strategy, and typically places an emphasis on education. Control measures used on the noxious plant species impacting this ecological site include chemical, biological, and cultural (burning, mowing, cutting) control methods. The best success has been found with an integrated pest management (IPM) strategy that incorporates one or several of these options along with education and prevention efforts (DiTomaso 2000)

### State 5 Conifer Encroached State

Rocky Mountain juniper (Juniperus scopulorum) and Douglas fir (Pseudotsuga menziesii) encroachment is common on the Droughty ecological site and is generally focused in areas with east and north facing slopes. Under the Reference State, conifers may exist; however, this is limited to two trees per acre on north and east facing slopes and is considered a trace canopy cover. Conifer encroachment likely occurs in the late stages of the Reference State and Altered State (see State-and-transition model) where there is an increase of bare ground and reduced vigor of mid-statured bunchgrasses due to a combination of factors. Fire suppression and improper grazing management are the two most common triggers. The exact conditions in which juniper begins to encroach vary, however the trend points to a combination of one or more of the following factors; moderately heavy to heavy grazing, reduced (non-existent) fire frequency, increased atmospheric carbon, and generally warmer climate (compared to that of pre-settlement). When heavy grazing occurs areas in the plant canopy open allowing for seed dispersal by bird or overland flow via rills on neighboring sites. The effects of conifer encroachment are not immediately noticed however over time as conifer canopy increases; light and water interception increase which reduces opportunities for herbaceous plants. Literature (Barrett, 2007) suggests that for precipitation to penetrate the juniper canopy, events must be greater than 0.30 inches. Increase juniper canopy creates perching sites for predators which reduces site suitability for greater sage grouse. Studies (Miller et al 2000) based in a similar community to the Conifer Encroached communities of this LRU suggest following a phased approach to characterize the stand. Not unlike the Western Juniper community discussed in Miller et al, these communities of Montana exhibit 3 different phases based, at this time, on qualitative information.

## Community 5.1 Conifer Community Phase I

Phase I (Early) is defined by actively expanding conifer cover with generally less than 5 percent canopy cover, where trees limbs touch the ground. This early stage has not lost its hydrologic functions; however, herbaceous plant communities may show signs of reduced production and species richness. Control methods include mechanical removal and prescribed fire. Prescribed fire is still effective management in this phase as it still contains the necessary native plants for recovery. The tree canopy density is also low enough that risk of a dangerously hot fire is reduced.

### Community 5.2 Conifer Community Phase II

Phase II (Mid Stage) is still actively expanding. Canopy cover may reach up to 15 percent. Due to the more mature trees present, seed production is very high. This Mid Phase begins to highly restrict herbaceous and shrub cover. Conifers tend to be codominant with herbaceous plants. Hydrology is departing from reference with rills becoming longer and, in isolated areas, erosional gullies may begin to form. Control methods of the Mid Stage should focus on mechanical treatment (chainsaw cutting or mastication) as there is a high risk of catastrophic and potentially hot, sterilizing fire. Post treatments of slash are currently being evaluated but typically include chipping, lop-and-scatter,

and slash piling. Lop-and-scatter and slash pile treatments are often burned to remove excess woody debris. Qualitative assessments of these treatments suggests native bunchgrasses respond immediately in this stage and post treatment grazing management including two growing seasons of rest may be necessary while plants increase their basal area.

## Community 5.3 Conifer Community Phase III

Phase III (Late stage) has juniper cover exceeding 20 percent and has slowed to resemble a forest condition. Lower limbs of trees begin to die and the rangeland shrub cover is nearly lost. Travel through this community is increasingly difficult. Conifers become dominant with herbaceous production greatly decreased. Bare ground increases and hydrologic function is nearly lost compared to a grass/shrub community. Late Stage Phase should focus more on restoration than control because the necessary plants are not present to cross the threshold back to the Reference State. Site stability and hydrologic function are lacking in this phase, so mechanical removal of conifer will be necessary and prescribed fire is not suggested. Post treatment rest from grazing is important as remaining native grasses will be reduced in stature and will be susceptible to accidental overgrazing as livestock seek out these tender plants. After a rest period, light and brief grazing may help initiate tillering of bunchgrasses.

### Pathway 5.1A Community 5.1 to 5.2

Over time Phase I community expands to increase in both height and width. The driver for this pathway is primarily lack of fire; however, heavy grazing (utilization that exceeds 50 percent) can help reduce herbaceous competition and expose soil for seed contact. Increase atmospheric carbon dioxide has also been shown to increase coniferous tree and shrub growth (Archer et al. 2017).

### Pathway 5.2B Community 5.2 to 5.3

Over time Phase II community expands to increase in both height and width. The driver for this pathway is primarily lack of fire however heavy grazing (utilization that exceeds 50 percent) can help reduce herbaceous competition and expose soil for seed contact. Increase atmospheric carbon dioxide has also been shown to increase coniferous tree and shrub growth (Archer et al. 2017).

## Transition T1B State 1 to 2



The Reference State (1) transitions to the Altered State (2) if bluebunch wheatgrass decreases to 15 percent, by dry weight, or if bare ground cover increases by 10 percent. The driver for this transition is the loss of taller bunchgrasses, which creates open areas in the plant canopy with bare soil. Soil erosion results in decreased soil fertility, driving transitions to the Altered State. There are several other key factors signaling the approach of transition T1B: increases in soil physical crusting, decreases in cover of cryptogamic crusts, decreases in soil surface aggregate stability or evidence of erosion including water flow patterns, development of plant pedestals, and litter movement. The trigger for this transition is improper grazing management and long-term drought leading to a decrease in bluebunch wheatgrass composition to 15 percent and reduction in total plant canopy cover.

## Transition T1C State 1 to 3

The Reference State (1) transitions to the Degraded State (3) when bluebunch wheatgrass is completely removed from the plant community. Idaho fescue and needle and thread are subdominant to short-statured bunchgrasses

such as Sandberg and Cusick's bluegrass. The trigger for this transition is loss of most mid-statured bunchgrasses, which creates open spaces with bare soil. Soil erosion as a result of increased bare ground and shallow roots decreases soil fertility, driving transitions to the Degraded State. There are several other key factors signaling the approach of transition T1C: increases in soil physical crusting, decreases in cover of cryptogamic crusts, decreases in soil surface aggregate stability or evidence of erosion including rills, water flow patterns, development of plant pedestals, and litter movement. The driver for this transition is improper grazing management, intense or repeated fires, or heavy human disturbance. Rapid transition is generally realized where livestock are confined to small pastures for long periods of time such as feeding areas, horse pastures, and bull lots. Degradation may be so extreme that traditional restoration methods may not be successful and require extensive mechanical and financial inputs.

## Transition T1D State 1 to 4

Healthy plant communities are most resistant to invasion; however, regardless of grazing management, without some form of active weed management (chemical, mechanical, or biological control), the Reference State (1) can transition to the Invaded State (4) in the presence of aggressive invasive species such as spotted knapweed, leafy spurge, and cheatgrass. The Central Rocky Mountains tend to resists invasion; however, recent dry climate cycles, repeated heavy grazing or intense human activities can open the interspaces of the bunchgrass community and allow for encroachment. Long-term stress conditions for native species (e.g., overgrazing, drought, and fire) accelerate this transition. If populations of invasive species reach critical levels, the site transitions to the Invaded State. The trigger for this transition is the presence of aggressive invasive species with invasive species composition by dry weight approaching 10 percent.

## Transition T1E State 1 to 5

Conifer tree/shrub count exceeds two stems per acre. The trigger of crossing a threshold is the presence of seeds and/or other viable material of these tree species.

## Restoration pathway R2A State 2 to 1



The Altered State (2) has lost soil or vegetation attributes to the point that recovery to the Reference State (1) will require reclamation efforts such as soil rebuilding, intensive mechanical and cultural treatments, and revegetation. Examples of mechanical treatment may be brush control while cultural treatments may include prescribed grazing, targeted brush browsing, or prescribed burning. Grazing practices that promote bluebunch wheatgrass is primarily light to moderate grazing during the critical season (late June through July) or fall and dormant season of moderate; however, heavy utilization may not have negative impacts (Dormaar and Willms 1998) though grazing should match the species composition of the site prior to exceeding moderate utilization. Low intensity prescribed fires to reduce competitive increaser plants such as needle and thread and Sandberg bluegrass. A low intensity fire will also reduce big sagebrush densities. In areas with potential of annual grass infestation, fire should be carefully planned or avoided. The drivers for this restoration pathway are reclamation efforts along with proper grazing management.

#### **Conservation practices**

Brush Management
Prescribed Burning
Fence
Livestock Pipeline

Prescribed Grazing
Grazing Land Mechanical Treatment
Range Planting

## Transition T2A State 2 to 3

As improper grazing continues vigor of bunchgrasses will decrease, and the shorter grasses and shrubs will increase towards the Degraded State (3). Improper grazing management for this state can be defined as grazing events that exceed moderate grazing (40 to 50 percent grazing use), grazing season that exceeds half of the growing season, and/or grazing events that consume the plant regrowth in the same growing season. Highly managed grazing events that exceed moderate grazing levels for short timeframes are generally not included in this definition due to increased rest periods between these grazing events. Additionally, prolonged drought will provide a competitive advantage to shrubs allowing them to become co-dominant with grasses. Shrub canopy will increase however shrubs will express low growth forms. Mat forming forbs will also increase. Key transition factors: increase of native shrub canopy cover; reduction in bunchgrass production; decrease in total plant canopy cover and production; increases in mean bare patch size; increases in soil crusting; decreases in cover of cryptobiotic crusts; decreases in soil aggregate stability; and/or evidence of erosion including water flow patterns and litter movement.

## Transition T2B State 2 to 4

Invasive species can occupy the Altered State (2) and drive it to the Invaded State (4). The Altered State is at risk if invasive seeds and other viable material are present. The driver for this transition is more than 10 percent dry weight of invasive species. The trigger is the presence of seeds and other viable material of invasive species.

## Transition T2D State 2 to 5

Conifer tree/shrub count exceeds two stems per acre. The trigger of crossing a threshold is the presence of seeds and other viable material of these tree species.

## Restoration pathway R3B State 3 to 1

The Degraded State (3) has lost soil or vegetation attributes to the point that recovery to the Reference State (1) will require reclamation efforts, such as soil rebuilding, intensive mechanical treatments, and revegetation. Studies suggest (Whitford et al 1989) a mulch with high carbon to nitrogen ratio such as wood chips or bark in low moisture scenarios can be beneficial for slow mobilization of plant available nitrogen. Biochar may also be added to the system to improve Soil Organic Carbon (SOC) which should improve Cation Exchange Capacity (CEC), microbial activity, and hydrologic conductivity (Stavi 2012). The drivers for the restoration pathway are removal of increaser species, restoration of native bunchgrass species, persistent management of invasives and shrubs, and proper grazing management. Without continued control, invasive and shrub species are likely to return (probably rapidly) due to presence of seeds and/or other viable material in the soil and management related increases soil disturbance.

### **Conservation practices**

**Brush Management** 

**Prescribed Burning** 

## Restoration pathway R3A State 3 to 2

Since the deep-rooted bunchgrass plant community has been removed, restoration to the Altered State (2) is

unlikely unless a seed source is available. If a sufficient amount of bunchgrass remains on the site, chemical application or biological control in conjunction with proper grazing management, can reduce the amount of shrubs and invasive species and restore the site to the Shortgrass Community (2.2). Grazing management strategies that follow light grazing and allow for long periods of rest will allow for limited recovery of remaining bunchgrasses, however grazing management alone may not directly result in restoration. Restoration methods such as reseeding may be necessary Low intensity fire can be utilized to reduce shrub competition and allow the reestablishment of grass species. Caution must be used when considering fire as a management tool on sites with fire tolerant shrubs such as rubber rabbitbrush, as these shrubs will re-sprout after a burn. Broom snakeweed and fringed sagewort may or may not re-sprout depending on conditions (USDA Forest Service 2011).

## Transition T3A State 3 to 4

Invasive species can occupy the Degraded State (3) and drive it to the Invaded State (4). The Degraded State is at risk of this transition occurring if invasive seeds or viable material are present. The driver for this transition is presence of critical population levels of invasive species. The trigger is the presence of seeds or viable material of invasive species. This state has sufficient bare ground that the transition could occur simply due to presence or introduction of invasive seeds or viable material. This is particularly true of aggressive invasive species such as spotted knapweed and cheatgrass. This transition could be assisted by overgrazing (failure to adjust stocking rate to declining forage production), long-term lack of fire, or extensive drought.

## Transition T3C State 3 to 5

Conifer tree/shrub count exceeds two stems per acre. The trigger of crossing a threshold is the presence of seeds and other viable material of these tree species.

## Restoration pathway R4C State 4 to 1

Restoration of the Invaded State (4) to the Reference State (1) requires substantial energy input. The drivers for the restoration pathway are removal of invasive species, restoration of native bunchgrass species, persistent management of invasive species, and proper grazing management. Without continued control, invasive species are likely to return (probably rapidly) due to the presence of seeds and other viable material in the soil and management related practices that increase soil disturbance. If invaded by conifer encroachment, treatment depends on the condition of the rangeland. See Plant community 4.1 for alternative measures of restoration. Sites that have transitioned from the Degraded State (3) to the Invaded State (4) may be severely lacking soil and vegetative properties that will allow for restoration to the Reference State. Hydrologic function damage may be irreversible especially with accelerated gully erosion.

## Restoration pathway R4B State 4 to 2

If invasive species are removed before remnant populations of bunchgrasses have been drastically reduced the Invaded State (4) can return to the Altered State. The driver for the reclamation pathway is weed management with possible reseeding. Continued Integrated Pest Management (IPM) will be required as many of the invasive species that can occupy the Invaded State have extended dormant seed life. The trigger is invasive species control.

## Restoration pathway R4A State 4 to 3

If invasive species are removed the site could return to the Degraded State (3). Without sufficient remnant populations of preferred plants the Invaded State (4) is not likely to return to any of the other states. The driver for the reclamation pathway is weed management without reseeding. The trigger is invasive species control. The invading species cause a significant increased soil loss due to lack of ground cover (Lacey et al. 1989).

#### **Transition T4B**

#### State 4 to 5

Canopy cover of conifer tree/shrub cover exceeds two stems per acre. The threshold change is triggered by the presence of seeds and other viable material of invasive species.

### Restoration pathway R5A State 5 to 1

Depending on the level of conifer canopy cover and its impact on rangeland health, restoration efforts may be simply focus on removal of coniferous trees and shrubs to restore the Conifer Encroached State (5) to the Reference State (1). If utilizing the phases established by Miller et al., management and restoration methods will vary. A large majority of the conifer encroachment in MLRA 43B will fall into the early two phases described by Miller et al. Phase I may exhibit None-Slight to Moderate departures from rangeland health where removal of the conifers via Brush Management or Prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment or rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is encouraged. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs because prescribed fire is not a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw waddles as well as range planting or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool-season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since herbaceous plants will likely have been depleted under a Phase III encroachment, there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

### Restoration pathway R5B State 5 to 2

The Conifer Encroached State (5) Phases I and II will often resemble early stages of the Mixed Sagebrush Community (2.1) of the Altered State (2) on this site. If utilizing the phases established by Miller et al., management and restoration methods will vary. A large majority of the conifer encroachment in MLRA 43B will fall into the early two phases described by Miller et al. Phase I may exhibit None-Slight to Moderate departures from rangeland health where removal of the conifers via Brush Management or Prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment and rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is encouraged. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs with prescribed fire not being a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw waddles as well as range planting or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site

are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since herbaceous plants will likely have been depleted under a Phase III encroachment, there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

## Restoration pathway R5C State 5 to 3

The Conifer Encroached State (5) Phases II and III will likely resemble the Degraded State (3) due to reduced midstatured bunchgrasses. If utilizing the phases established by Miller et al., management and restoration methods will vary. A large majority of the conifer encroachment in MLRA 43B will fall into the early two phases described by Miller et al. This restoration pathway is exceedingly rare as it is typically not cost effective for land managers to manage for a degraded state. Phase I may exhibit None-Slight to Moderate departures from rangeland health where removal of the conifers via brush management or prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment or rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is encouraged. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs with prescribed fire not being a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw waddles as well as range planting or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since herbaceous plants will likely have been depleted under a Phase III encroachment, there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting and/or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

### Restoration pathway R5D State 5 to 4

If utilizing the phases established by Miller et al., management and restoration methods will vary. A large majority of the conifer encroachment in MLRA 43B will fall into the early two phases described by Miller et al. This restoration pathway is exceedingly rare as it is typically not cost effective for land managers to manage for the Invaded State. Phase I may exhibit None-Slight to Moderate departures from rangeland health where removal of the conifers via brush management or prescribed fire combined. If mechanical removal of conifers is utilized, no grazing management is needed assuming relatively conservative management had been used prior to treatment. If prescribed fire is utilized, short term grazing deferment or rest is suggested. Given a short time removal of a Phase I encroachment will recover to Reference. Proactive pest management is absolutely necessary with presence of invasive herbaceous species. Phase II Encroachment may require a more intense mechanical removal of trees/shrubs with prescribed fire not being a feasible method of control as this community may be at risk of catastrophic fire due to canopy density. Phase II displays a Moderate departure from Reference suggesting an overall instability of the site such as reduced herbaceous production, reduced functional/structural groups (e.g. reduced mid-statured bunchgrasses), increase rill frequency and length, and possibly increased bare ground. Increased post treatment grazing management may be necessary. Grazing management may be as simple as short term growing season deferment however long term rest may be necessary in the latter stages of Phase II encroachment. Latter stages of Phase II encroachment will likely require some short term erosion mitigation such as straw waddles as well as range planting or critical area planting to re-establish any loss of native herbaceous plants particularly mid-statured cool season bunchgrasses. Phase III Encroachment canopy cover resembles forested sites with larger trees and shrubs. Forest management style tree removal (woody debris and logs removed

from the site) will be necessary prior to any prescribed burning as to prevent the fire from burning too hot. The result of a prescribed fire on this site are typically unknown as seed sources of native herbaceous plants are usually limited to small patches. Since herbaceous plants will likely have been depleted under a Phase III encroachment, there is an opportunity for large areas of bare ground, increase rill and in some cases gully erosion. Post treatment will require range planting and/or critical area seeding, erosion control, pest management, and possibly soil carbon amendments (biochar). Grazing management (primarily rest) will be necessary to ensure any new seedling establishment.

### Additional community tables

Table 8. Community 1.1 plant community composition

Group	Common Name	Symbol	Scientific Name	Annual Production (Kg/Hectare)	Foliar Cover (%)
Grass	/Grasslike	-!	<del> </del>		
1	Midstatured Bunchgras	sses		970–1171	
	bluebunch wheatgrass	PSSP6	Pseudoroegneria spicata	448–583	25–35
	green needlegrass	NAVI4	Nassella viridula	67–202	1–5
	basin wildrye	LECI4	Leymus cinereus	0–112	0–5
	Columbia needlegrass	ACNE9	Achnatherum nelsonii	11–45	3–5
	spike fescue	LEKI2	Leucopoa kingii	0–22	0–2
	rough fescue	FECA4	Festuca campestris	0–6	0
2	Increaser Bunchgrasse	es and sedg	ges	168–241	
	Idaho fescue	FEID	Festuca idahoensis	112–202	5–10
	needle and thread	HECO26	Hesperostipa comata	0–67	1–5
	Sandberg bluegrass	POSE	Poa secunda	11–67	0–3
	sedge	CAREX	Carex	0–67	0–3
	prairie Junegrass	KOMA	Koeleria macrantha	22–45	1–2
	Cusick's bluegrass	POCU3	Poa cusickii	0–45	0–2
	needleleaf sedge	CADU6	Carex duriuscula	11–22	0–1
	timber oatgrass	DAIN	Danthonia intermedia	0–22	0–1
	Porter brome	BRPO2	Bromus porteri	0–17	0–1
3	Rhizomatous Grasses			78–129	
	thickspike wheatgrass	ELLA3	Elymus lanceolatus	45–157	0–2
	western wheatgrass	PASM	Pascopyrum smithii	0–84	0–2
	plains reedgrass	CAMO	Calamagrostis montanensis	0–45	0–1
Forb		-!	-	•	
4	Forbs			224–336	
	lupine	LUPIN	Lupinus	45–135	0–2
	American vetch	VIAM	Vicia americana	45–112	0–2
	sulphur-flower buckwheat	ERUM	Eriogonum umbellatum	0–90	0–1
	common yarrow	ACMI2	Achillea millefolium	22–67	0–1
	western stoneseed	LIRU4	Lithospermum ruderale	22–67	0–1
	goldenbanner	THERM	Thermopsis	0–45	0–1
	beardtongue	PENST	Penstemon	22–45	0–1
	spiny phlox	РННО	Phlox hoodii	0–22	0–1
	rosy pussytoes	ANRO2	Antennaria rosea	0–22	0–1

	geranium	GERAN	Geranium	0–17	0–1
	false dandelion	NOTHO5	Nothocalais	0–17	0–1
	cinquefoil	POTEN	Potentilla	0–17	0–1
	Indian paintbrush	CASTI2	Castilleja	0–11	0–1
	northern bedstraw	GABO2	Galium boreale	0–11	0–1
	balsamroot	BALSA	Balsamorhiza	0–11	0
	elkweed	FRSP	Frasera speciosa	0–6	0
Shrub	/Vine	-	-	•	
5	Shrubs		224–308		
	mountain big sagebrush	ARTRV	Artemisia tridentata ssp. vaseyana	6–185	5–15
	threetip sagebrush	ARTR4	Artemisia tripartita	0–67	0–2
	spineless horsebrush	TECA2	Tetradymia canescens	0–56	0–2
	common snowberry	SYAL	Symphoricarpos albus	6–45	0–1
	rubber rabbitbrush	ERNA10	Ericameria nauseosa	0–45	0–1
	shrubby cinquefoil	DAFR6	Dasiphora fruticosa	0–45	0–1
	yellow rabbitbrush	CHVI8	Chrysothamnus viscidiflorus	0–28	0–1
	broom snakeweed	GUSA2	Gutierrezia sarothrae	0–22	0–1
	Woods' rose	ROWO	Rosa woodsii	0–22	0–1
	chokecherry	PRVI	Prunus virginiana	0–22	0–1
Tree		-			
6	Trees			0–22	
	Douglas-fir	PSME	Pseudotsuga menziesii	0–11	0
	Rocky Mountain juniper	JUSC2	Juniperus scopulorum	0–11	0
	quaking aspen	POTR5	Populus tremuloides	0–11	0

### **Animal community**

The Droughty Ecological Site provides for a variety of wildlife habitat for an array of species. Prior to the settlement of this area, large herds of antelope, elk and bison roamed. Though the bison have been replaced, mostly with domesticated livestock; elk and antelope still frequently utilize this largely intact landscape for yearlong habitat in areas adjacent to forest.

The relatively high grass component of the Reference Community provides excellent nesting cover for multiple neotropical migratory birds that select for open grasslands such as the Long-billed curlew and McCown's longspur.

Greater sage grouse may be present on sites with suitable habitat, typically requiring a minimum of 15 percent sagebrush canopy cover (Braun et al. 1977). The Reference State (1) is likely to have this minimum sagebrush cover for sage grouse presence given its moderate sagebrush canopy cover. Also, the potentially diverse forb component of the Reference State may provide the important early season (spring) foraging habitat for the Greater sage grouse. Other communities on the site with sufficient sagebrush cover may harbor sage grouse populations specifically Communities 1.3 and 2.1 where big sagebrush populations increased under a reduced fire regime. Also as sagebrush canopy cover increases under Altered States and, to a limited extent, the Degraded State 3.1; Pygmy rabbit, Brewer's sparrow, pronghorn antelope, elk, and mule deer use may also increase.

Managed livestock grazing is suitable on this site due to the potential to produce an abundance of high quality forage. This is often a preferred site for grazing by livestock, and animals tend to congregate in these areas. In order to maintain the productivity of the site, grazing on adjoining sites with less production must be managed carefully to be sure utilization on this site is not excessive. Management objectives should include maintenance or improvement of the native plant community. Careful management of timing and duration of grazing is important.

Shorter grazing periods and adequate deferment during the growing season are recommended for plant maintenance, health, and recovery. According to McLean and Wikeen 1985, early season defoliation of bluebunch wheatgrass can result in high mortality and reduced vigor of plants. They also suggest, based on prior studies, that regrowth is necessary before dormancy to reduce injury of bluebunch wheatgrass.

Grazing season has more influence on winterfat than grazing intensity. Late winter or early spring grazing is detrimental. However, early winter grazing may actually be beneficial (Blaisdell and Holmgren 1984).

Continual, non-prescribed grazing of this site will be detrimental, will alter the plant composition and production over time, and will result in transition to the Altered State (2). Transition to other states will depend on duration of poorly managed grazing as well as other circumstances such as weather conditions and fire frequency.

The Altered State (2) is subject to further degradation to the Degraded State (3) or Invaded State (4). Management should focus on grazing management strategies that will prevent further degradation, such as seasonal grazing deferment or winter grazing where feasible. Communities within this state are still stable and healthy under proper management. Forage quantity and/or quality may be substantially decreased from the Reference State (1).

Grazing is possible in the Invaded State (4). Invasive species are generally less palatable than native grasses. Forage production is typically greatly reduced in this state. Due to the aggressive nature of invasive species, sites in the Invaded State (4) face increased risk for further degradation to the invasive dominated community. Grazing has to be carefully managed to avoid further soil loss and degradation and possible livestock health issues.

Prescriptive grazing can be used to manage invasive species. In some instances, carefully targeted grazing (sometimes in combination with other treatments) can reduce or maintain species composition of invasive species. In the Degraded State, grazing may be possible but is generally not economically and/or environmentally sustainable.

### **Hydrological functions**

The hydrologic cycle functions best in the Reference State (1) with good infiltration and deep percolation of rainfall; however, the cycle degrades as the vegetation community declines. Rapid rainfall infiltration, high soil organic matter, good soil structure, and good porosity accompany high bunchgrass canopy cover (Thurow et al. 1986). High ground cover reduces rain drop impact on the soil surface, which keeps erosion very low.

Water leaving the site will have minimal sediment load, which allows for high water quality in associated streams. High rates of infiltration will allow water to move below the rooting zone during periods of heavy rainfall. The Bluebunch Community (1.1) should have no rills or gullies present and drainage ways should be vegetated and stable. Water flow patterns will not be present under the Reference State (1). Plant pedestals are essentially non-existent. Plant litter remains in place and is not moved by wind or water.

Improper grazing management results in a community shift to the Mixed Bunchgrass Community (1.2). This plant community has a similar canopy cover, but bare ground will be less than 15 percent. Therefore, the hydrologic cycle is functioning at a level similar to the water cycle in the Bluebunch Bunchgrass Community (1.1). When compared to the Bluebunch Bunchgrass Community (1.1) the infiltration rates of Mixed Bunchgrass Community (1.2) are slightly reduced and surface runoff is slightly higher though often minimal.

In the Shortgrass-Shrub Community (2.2), Degraded State (3) and the Invaded State (4) canopy and ground cover are greatly reduced compared to the Reference State (1), which impedes the hydrologic cycle. Infiltration will decrease and runoff will increase due to reduced ground cover, presence of shallow-rooted species, rainfall splash, soil capping, reduced organic matter, and weak soil structure. Sparse ground cover and decreased infiltration can combine to increase frequency and severity of flooding within a watershed. Soil erosion is accelerated, quality of surface runoff is poor, and sedimentation increases (McCalla et al. 1984).

Hydrology of the Conifer Encroached State (5) is highly variable however studies suggest that increased tree canopy affects interception of rainfall and reduces available soil moisture for herbaceous vegetation. This can negatively affect infiltration and increase runoff (Pierson et al. 2010)(Miller et al. 2000).

#### Recreational uses

This site provides recreational opportunities for hiking, horseback riding, big game and upland bird hunting. Some forbs have flowers that appeal to photographers. This site provides valuable open space.

#### **Wood products**

All stages of the Conifer Encroached State may express some wood products such as firewood, thin post and pole, and decorative wood.

### Other products

none

### Inventory data references

Information presented was derived from NRCS inventory data, literature, field observations, and personal contacts with range-trained personnel (i.e., used professional opinion of agency specialists, observations of land managers, and outside scientists).

#### References

- Arno, S.F. and G.E. Gruell. 1982. Fire History at the Forest-Grassland Ecotone in Southwestern Montana. Journal of Range Management 36:332–336.
- Bestelmeyer, B., J.R. Brown, K.M. Havstad, B. Alexander, G. Chavez, and J.E. Herrick. 2003. Development and Use of State and Transition Models for Rangelands. Jornal of Range Management 56:114–126.
- Bestelmeyer, B. and J.R. Brown. 2005. State-and-Transition Models 101: a Fresh Look at Vegetation Change.
- Bestelmeyer, B. and J. Brown. 2005. State-and-Transition Models 101: A Fresh look at vegetation change.
- Bestelmeyer, B.T., K. Moseley, P.L. Shaver, H. Sanchez, D.D. Briske, and M.E. Fernandez-Gimenez. 2010. Practical guidance for developing state-and-transition models. Rangelands 32:23–30.
- Bestelmeyer, B.T., J.C. Williamson, C.J. Talbot, G.W. Cates, M.C. Duniway, and J.R. Brown. 2016. Improving the Effectiveness of Ecological Site Descriptions: General State-and-Transition Models and the Ecosystem Dynamics Interpretive Tool (EDIT). Rangelands 38:329–335.
- Blaisdell, J.P. and R.C. Holmgren. 1984. Managing Intermountain Rangelands--Salt-Desert Shrub Ranges. General Tech Report INT-163. USDA Forest Service, Intermountain Forest and Range Experiment Station, Ogden, UT. 52.
- Braun, C.E., T. Britt, and R.O. Wallestad. 1977. Guidelines for Maintenance of Sage Grouse Habitats. Wildlife Society Bulletin (1973-2006) 5:99–106.
- Briske, D.D., B.T. Bestelmeyer, T.K. Stringham, and P.L. Shaver. 2008. Recommendations for Development of Resilience-Based State-and-Transition Models. Rangeland Ecology & Management 61:359–367.
- DiTomaso, J.M. 2000. Invasive weeds in Rangelands: Species, Impacts, and Management. Weed Science 48:255–

- Gruell, G.E. 1983. Fire and vegetative trends in the Northern Rockies: interpretations from 1871-1982 photographs. USDA Forest Service General Technical Report INT-158. U.S. Department of Agriculture, Forest Service, Intermountain Forest and Range Experiment Station, Ogden, UT. 1–117.
- Hayden, F.V. 1873. Sixth Annual Report of the United States Geological Survey of the Territories, embracing portions of Montana, Idaho, Wyoming, and Utah, being a report of progress of the explorations for the year 1872.
- Hobbs, J.R. and S.E. Humphries. 1995. An integrated approach to the ecology and management of plant invasions. Conservation Biology 9:761–770.
- Lesica, P. and S.V. Cooper. 1997. Presettlement vegetation of Southern Beaverhead County, MT.
- Lesica, P., S.V. Cooper, and G. Kudray. 2007. Recovery of Big Sagebrush Following Fire in Southwest Montana. Rangeland Ecology and Management 60:261–269.
- Masters, R. and R. Sheley. 2001. Principles and practices for managing rangeland invasive plants. Journal of Range Management 38:21–26.
- McCalla, G.R., W.H. Blackburn, and L.B. Merrill. 1984. Effects of Livestock Grazing on Infiltration Rates of the Edwards Plateau of Texas. Journal of Range Management 37:265–269.
- McLean, A. and S. Wikeem. 1985. Influence of season and intensity of defoliation on bluebunch wheatgrass survival and vigor in southern British Columbia. Journal of Range Management 38:21–26.
- Miller, R.F., T.J. Svejcar, and J.A. Rose. 2000. Impacts of western juniper on plant community composition and structure. Journal of Range Management 53:574–585.
- Moulton, G.E. and T.W. Dunlay. 1988. The Journals of the Lewis and Clark Expedition. Pages in University of Nebraska Press.
- Ross, R.L., E.P. Murray, and J.G. Haigh. July 1973. Soil and Vegetation of Near-pristine sites in Montana.
- Sturm, J.J. 1954. A study of a relict area in Northern Montana. University of Wyoming, Laramie 37.
- Thurow, T.L., W.H. Blackburn, and C.A. Taylor. 1986. Hydrologic Characteristics of Vegetation Types as Affected by Livestock Grazing Systems, Edwards Plateau, Texas. Journal of Range Management 39:505–509.
- Wilson, A.M., G.A. Harris, and D.H. Gates. 1966. Cumulative Effects of Clipping on Yield of Bluebunch wheatgrass. Journal of Range Management 19:90–91.

#### **Contributors**

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### **Approval**

Kirt Walstad, 3/04/2024

### Rangeland health reference sheet

Interpreting Indicators of Rangeland Health is a qualitative assessment protocol used to determine ecosystem condition based on benchmark characteristics described in the Reference Sheet. A suite of 17 (or more) indicators are typically considered in an assessment. The ecological site(s) representative of an assessment location must be known prior to applying the protocol and must be verified based on soils and climate. Current plant community cannot be used to identify the ecological site.

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Date	03/04/2021
Approved by	Kirt Walstad
Approval date	
Composition (Indicators 10 and 12) based on	Annual Production

Co	mposition (Indicators 10 and 12) based on Annual Production
Inc	licators
1.	Number and extent of rills: Rills will not exist under Reference State.
2.	Presence of water flow patterns: Water flow patters will not exist under Reference State.
3.	Number and height of erosional pedestals or terracettes: Pedestals and terracettes will not exist under Reference State.
4.	Bare ground from Ecological Site Description or other studies (rock, litter, lichen, moss, plant canopy are not bare ground): Bare ground will be less than 10 percent.
5.	Number of gullies and erosion associated with gullies: Gullies will not be present.
6.	Extent of wind scoured, blowouts and/or depositional areas: Wind scoured, blowouts, or depositional areas will not exist on this site under Reference State.
7.	Amount of litter movement (describe size and distance expected to travel): Litter movement is very limited on this site with herbaceous litter moving no more than 4 inches (10 centimeters) from its origin.

8. Soil surface (top few mm) resistance to erosion (stability values are averages - most sites will show a range of

	values): Soils of this site are stable and should have Ratings of 3-6 using the Soil Stability Methods.
	Soil surface structure and SOM content (include type of structure and A-horizon color and thickness): Soil structure at the surface is typically strong to medium fine granular. The A horizon should be 5-7 inches thick with color, when wet, typically ranging in Value of 3 or less and Chroma of 3 or less. Local geology may affect color, so it is important to reference the Official Series Description (OSD) for characteristic range
	Effect of community phase composition (relative proportion of different functional groups) and spatial distribution on infiltration and runoff: Evenly distributed across the site, bunchgrasses improve infiltration while rhizomatous grasses protect the surface from runoff forces. The Droughty ecological site is well drained and has a moderate infiltration rate. An even distribution of grasses with totaling greater than 70 percent of site production, approximately 5 percent cool season rhizomatous grasses, 10 percent shortgrass and 55 percent deep rooted midstatured bunchgrass. Forbs (15 percent), shrubs (15 percent), and trees (trace) account for the remaining percentage by weight.
	Presence and thickness of compaction layer (usually none; describe soil profile features which may be mistaken for compaction on this site): Not present, some soils profiles may contain an abrupt transition to an argillic horizon which can be interpreted as compaction; however, the soil structure is typically subangular blocky whereas a compaction layer will tend to be structureless.
	Functional/Structural Groups (list in order of descending dominance by above-ground annual-production or live foliar cover using symbols: >>, >, = to indicate much greater than, greater than, and equal to):
	Dominant: Cool season mid-statured bunchgrasses (bluebunch wheatgrass, Columbia needlegrass, Richardsons needlegrass, green needlegrass, spike fescue, rough fescue)
	Sub-dominant: forbs = shrubs > cool-season short-grasses > rhizomatous grasses >> trees
	Other:
	Additional:
3.	Amount of plant mortality and decadence (include which functional groups are expected to show mortality or decadence): Mortality in herbaceous species is not evident. Species with bunch growth forms may have some natural mortality in centers.
4.	Average percent litter cover (%) and depth ( in): Total litter cover ranges from 30-40 percent. Most litter is irregularly distributed on the soil surface and is typically less than 0.5 inches thick
	Expected annual annual-production (this is TOTAL above-ground annual-production, not just forage annual-production): Average production is 1800 pounds per acre (lb/ac) or 2017 kilograms per bectare (kg/ba)

Low production is 1340 lb/ac or 1502kg/ha High production is 2620 lbs/ac or 2937kg/ha

Production values can vary greatly due to site conditions and precipitation patterns.	These values represent the Relative
Value (RV) ranges for this siteoutliers within the Reference State do exist	

16. Potential invasive (including noxious) species (native and non-native). List species which BOTH characterize degraded states and have the potential to become a dominant or co-dominant species on the ecological site if their future establishment and growth is not actively controlled by management interventions. Species that become dominant for only one to several years (e.g., short-term response to drought or wildfire) are not invasive plants. Note that unlike other indicators, we are describing what is NOT expected in the reference state for the ecological site: dandelion (Taraxicum spp), cheatgrass (Bromus techtorum), field brome (Bromus arvensis), spotted knapweed (Centaurea stoebe), yellow toadflax (Linaria vulgaris), leafy spurge (Euphorbia esula), and Kentucky bluegrass (Poa pratensis)

Note: this list may not be fully comprehensive as unknown populations of weeds may exist

Native species with the ability to indicate degradation however species presence alone does not imply degradation: Sandberg bluegrass (*Poa secunda*), big sagebrush (*Artemisia tridentata*), three-tip sagebrush (*Artemisia tripartita*), broom snakeweed (Gutierrezia sarothrae), rubber rabbitbrush (Ericameria nauseosa), yellow rabbitbrush (Chrysothamnus viscidiflorus), Rocky Mountain juniper (*Juniperus scopulorum*), Douglas fir (*Pseudotsuga menziesii*)

17. **Perennial plant reproductive capability:** In the reference condition, all plants are vigorous enough for reproduction either by seed or rhizomes in order to balance natural mortality with species recruitment. Density of plants indicates that plants reproduce at levels sufficient to fill available resource.