

# Ecological site EX044B01A165 Thin Claypan (TCp) LRU 01 Subset A

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#### General information

**Provisional**. A provisional ecological site description has undergone quality control and quality assurance review. It contains a working state and transition model and enough information to identify the ecological site.

### **MLRA** notes

Major Land Resource Area (MLRA): 044B-Central Rocky Mountain Valleys

Major Land Resource Area (MLRA) 44B, Central Rocky Mountain Valleys, is nearly 3.7 million acres of southwest Montana and borders two MLRAs: 43B Central Rocky Mountains and Foothills and 46 Northern and Central Rocky Mountain Foothills.

The major watersheds of this MLRA are those of the Missouri and Yellowstone Rivers and their associated headwaters such as the Beaverhead, Big Hole, Jefferson, Ruby, Madison, Gallatin, and Shields Rivers. These waters allow for extensive irrigation for crop production in an area that would generally only be compatible with rangeland and grazing. The Missouri River and its headwaters are behind several reservoirs that supply irrigation water, hydroelectric power, and municipal water. Limited portions of the MLRA are west of the Continental Divide along the Clark Fork River.

The primary land use of this MLRA is production agriculture (grazing, small grain production, and hay), but there is some limited mining. Urban development is high with large expanses of rangeland converted to subdivisions for a rapidly growing population.

The MLRA consists of one Land Resource Unit (LRU) and seven climate based LRU subsets. These subsets are based on a combination of Relative Effective Annual Precipitation (REAP) and frost free days. Each subset expresses a distinct set of plants that differentiate it from other LRU subsets. Annual precipitation ranges from a low of 9 inches to a high near 24 inches. The driest areas tend to be in the valley bottoms of southwest Montana in the rain shadow of the mountains. The wettest portions tend to be near the edge of the MLRA at the border with MLRA 43B. Frost free days also vary widely from less than 30 days in the Big Hole Valley to around 110 days in the warm valleys along the Yellowstone and Missouri Rivers.

The plant communities of the MRLA are highly variable, but the dominant community is a cool-season grass and shrub-steppe community. Warm-season grasses have an extremely limited extent in this MLRA. Most subspecies of big sagebrush are present, to some degree, across the MLRA.

#### LRU notes

MLRA 44B has one LRU that covers the entire MLRA. The LRU has been broken into seven climate subsets based on a combination of Relative Effective Annual Precipitation (REAP) and frost free days. Each combination of REAP and frost free days results in a common plant community that is shared across the subset. Each subset is giving a letter designation of A through F for sites that do not receive additional water and Y for sites that receive additional water.

LRU 01 Subset A has a REAP of nine to 14 inches (228.6-355.6mm) with a frost free days range of 70 to 110 days. This combination of REAP and frost free days results in a nearly treeless sagebrush steppe landscape.

The soil moisture regime is Ustic, dry that borders on Aridic and has a Frigid soil temperature regime.

### Classification relationships

Mueggler and Stewart. 1980. Grassland and Shrubland habitat types of Western Montana

- 1. Stipa comata/Bouteloua gracilis h.t.
- 2. Agropyron spicatum/Bouteloua gracilis h.t.

Montana Natural Heritage Program Vegetation Classification

1. Stipa comata - *Bouteloua gracilis* Herbaceous Vegetation
(STICOM – BOUGRA) Needle and thread/Blue grama

Natural Heritage Conservation Rank-G5 / S5

Edition / Author- 99-11-16 / S.V. Cooper,

EPA Ecoregions of Montana, Second Edition:

Level I: Northwestern Forested Mountains

Level II: Western Cordillera

Level III: Middle Rockies & Northern Great Plains

Level IV: Paradise Valley

**Townsend Basin** 

Dry Intermontane Sagebrush Valleys

Shield-Smith Valleys

National Hierarchical Framework of Ecological Units:

Domain: Dry

Division: M330 – Temperate Steppe Division – Mountain Provinces

Province: M332 – Middle Rocky Mountain Steppe – Coniferous Forest – Alpine Meadow

Section: M332D – Belt Mountains Section M332E – Beaverhead Mountains Section

Subsection: M332Ej – Southwest Montana Intermontane Basins and Valleys

M332Dk - Central Montana Broad Valleys

## **Ecological site concept**

The Thin Claypan ecological site is an upland site formed from alluvium and is on slopes less than 15 percent. The site does not receive additional moisture from a water table or flooding. The Thin Claypan site has less than four (4) inches of the mineral surface over a heavy clay layer. This clay layer has greater than 35 percent clay. The site is moderately deep to very deep and has no root-restrictive layers within 20 inches (50cm). The site has a natric or relic natric horizons and is not strongly or violently effervescent within four inches of the mineral surface. This natric or relic natric horizon has a columnar structure, abrupt water restrictive clay layer present within 8" of soil surface. The surface of the site has less than five percent stone and is not skeletal, with less than 35 percent rock fragments in the 10 to 20-inch depth. Calcium carbonates may increase with depth.

### **Associated sites**

EX044B01A001	Clayey (Cy) LRU 01 Subset A The clayey ecological site occupies adjacent sites, often the hills above the Thin Claypan site.
EX044B01A006	Claypan (Cp) LRU 01 Subset A The Claypan ecological site occupies the same landscape as the Thin Claypan and is frequently adjacent to or intermixed with Thin Claypan sites. Plant communities are very similar and will share similar state and transition models.

### Similar sites

## EX044B01A006 Claypan (Cp) LRU 01 Subset A

The Claypan ecological occupies the same landscape as the Thin Claypan and is frequently adjacent to or intermixed with Thin Claypan sites. Plant communities are very similar and will share similar state and transition models. Claypan differs from other ecological sites in that it has higher plant productivity and deeper soil over the natric horizon.

Table 1. Dominant plant species

Tree	Not specified
	<ul><li>(1) Artemisia tridentata ssp. wyomingensis</li><li>(2) Sarcobatus vermiculatus</li></ul>
Herbaceous	<ul><li>(1) Pseudoroegneria spicata</li><li>(2) Hesperostipa comata</li></ul>

## **Legacy ID**

R044BA165MT

## Physiographic features

The Thin Claypan site exists on gently sloping (less than 15 percent slope) low hills near the valley bottom or on fan remnants.

Table 2. Representative physiographic features

Landforms	<ul><li>(1) Intermontane basin &gt; Hill</li><li>(2) Intermontane basin &gt; Fan remnant</li></ul>
Flooding frequency	None
Ponding frequency	None
Elevation	4,500–6,500 ft
Slope	2–15%
Water table depth	40 in
Aspect	Aspect is not a significant factor

### **Climatic features**

The Central Rocky Mountain Valleys MLRA has a continental climate. Fifty to sixty percent of the annual long-term average precipitation falls between May and August. Average precipitation for LRU 01 Subset A is 12 inches (305mm), and the frost-free period averages 78 days. Precipitation is highest in May and June.

Table 3. Representative climatic features

Frost-free period (characteristic range)	70-110 days
Freeze-free period (characteristic range)	110-140 days
Precipitation total (characteristic range)	9-14 in
Frost-free period (actual range)	70-110 days
Freeze-free period (actual range)	110-140 days
Precipitation total (actual range)	9-14 in
Frost-free period (average)	78 days
Freeze-free period (average)	125 days
Precipitation total (average)	12 in

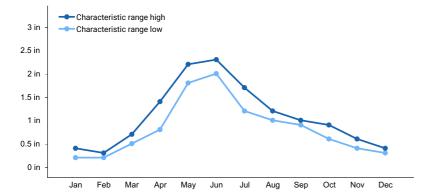


Figure 1. Monthly precipitation range

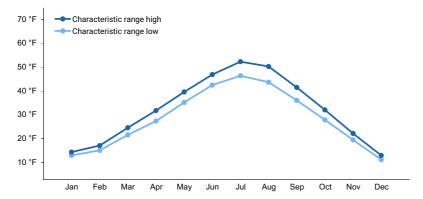


Figure 2. Monthly minimum temperature range

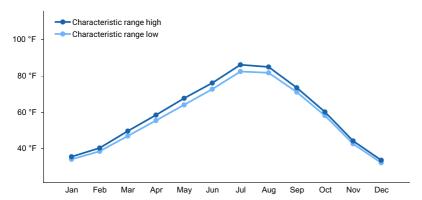


Figure 3. Monthly maximum temperature range

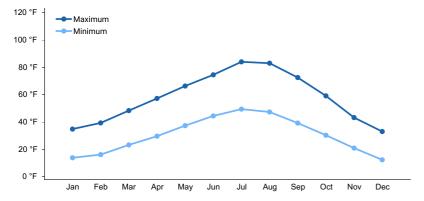


Figure 4. Monthly average minimum and maximum temperature

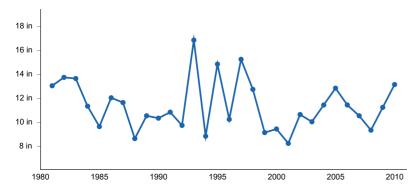


Figure 5. Annual precipitation pattern

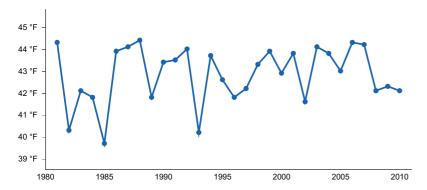


Figure 6. Annual average temperature pattern

### Climate stations used

- (1) HELENA RGNL AP [USW00024144], Helena, MT
- (2) GARDINER [USC00243378], Gardiner, MT
- (3) DILLON AP [USW00024138], Dillon, MT
- (4) DEER LODGE 3 W [USC00242275], Deer Lodge, MT
- (5) TRIDENT [USC00248363], Three Forks, MT
- (6) BOULDER [USC00241008], Boulder, MT
- (7) DILLION U OF MONTANA WESTERN [USC00242409], Dillon, MT
- (8) ENNIS [USC00242793], Ennis, MT
- (9) TWIN BRIDGES [USC00248430], Sheridan, MT
- (10) TOWNSEND [USC00248324], Townsend, MT

### Influencing water features

The site has a clay layer near the surface which slows infiltration. Runoff is high.

### Wetland description

This site is not associated with wetland characteristics.

### Soil features

This site exists on sedimentary-originated soil with a high sodium (natric or relic natric) horizon that has a columnar structure (within 20 inches of the soil surface) and an abrupt root or water-restrictive clay layer within 4 inches of the surface. As a result of this high clay layer, infiltration rates are slow. Clay content in the top four inches of soil is variable; however, an argillic horizon will have a significant increase in clay over the surface. Surface soil structure will range from fine granular to subangular blocky. An E horizon may be present, and if so, it will have a weak, platy structure. The structure of the clay layer will be columnar.

Common soil series include Dyce and Doolittle. As these soils may express a range of characteristics, it is necessary to use the most current ecological site key with an onsite soil pit to determine the correct ecological site.

Table 4. Representative soil features

Parent material	(1) Alluvium–sedimentary rock
Surface texture	(1) Clay loam (2) Loam
Family particle size	(1) Fine
Drainage class	Well drained
Permeability class	Moderately slow to slow
Surface fragment cover <=3"	0–10%
Surface fragment cover >3"	0–1%
Calcium carbonate equivalent (2-10in)	5–15%
Clay content (0-2in)	15–27%
Soil reaction (1:1 water) (2-10in)	7.4–8.6

## **Ecological dynamics**

The Thin Claypan Ecological Site occurs in a relatively small landscape; however, slight variations within the plant community may exist due to elevation, frost-free days, and relative effective annual precipitation.

The reference plant community is dominated by bluebunch wheatgrass (*Pseudoroegneria spicata*) and needle and thread. Subdominant species may include western wheatgrass, Wyoming big sagebrush (*Artemisia tridentata* ssp. wyomingensis), winterfat (*Krascheninnikovia lanata*), and dotted gayfeather (*Liatris punctata*). This potential is suggested by investigations showing a predominance of perennial grasses on near-pristine range sites (Ross et al., 1973). In the reference plant community, shrubs comprise 10 to 15 percent of the community, which tends to be higher than many other ecological sites in this MLRA.

The natural fire return interval of the area is highly variable, ranging up to 100 years, but due to the sparse vegetation and low production, natural fire is exceedingly rare.

A shift to the dominance of shrubs may occur in response to improper grazing management, drought, or where big sagebrush occurs due to a lack of fire. As the mid-stature bunchgrasses decrease, shrub encroachment by a variety of species, including broom snakeweed (*Gutierrezia sarothrae*), prairie sagewort (*Artemisia frigida*), Wyoming big sagebrush, yellow rabbitbrush (*Chrysothamnus viscidiflorus*), rubber rabbitbrush (*Ericameria nauseosa*), and plains prickly pear (*Opuntia polyacantha*), occurs within this site. Shrub dominance and grass loss are associated with soil erosion and, ultimately, thinning of the native soil surface. Subsequent loss of soil could lead to a Degraded Shortgrass State. All states could also lead to the Invaded State when there is a lack of weed prevention and control measures.

Historical records indicate that, prior to the introduction of livestock (cattle and sheep) during the late 1800s, elk and bison grazed this ecological site. Grazed areas received periodic high intensity, short duration grazing pressure due to bison's nomadic nature and herd structure. Livestock forage was noted as being minimal in areas recently grazed by bison (Lesica and Cooper 1997). Meriwether Lewis documented that he was met by 60 Shoshone warriors on horseback in August 1805, and the Corps of Discovery was later supplied with horses by the same band of Shoshone. This suggests that the areas near the modern-day Montana towns of Twin Bridges, Dillon, Grant, and Dell were grazed by an untold number of horses for nearly 50 years prior to the large introduction of cattle and sheep. The gold boom of the 1860s brought the first herds of livestock overland from Texas, and homesteaders began settling the area. During this time, cattle were the primary domestic grazers in the area. In the 1890s, sheep production increased by more than 400 percent and dominated the livestock industry until the 1930s. Since then, cattle production has dominated the region's livestock industry (Wyckoff and Hansen 2001).

Due to the alkaline pH and high clay content of the soils on this site, the potential for dryland farming is low, and it is

rarely successful over the long term.

Dense clubmoss (*Selaginella densa*), in general, is a minor component of the reference plant community of this ecological site. The conditions that created large cover classes of clubmoss on this site point to a history of continuous (year-long) or moderate spring grazing use (Sturm 1954). In some situations, the site could be old crop fields that have reverted back to rangeland. In this case, clubmoss is helping reduce erosion and increase site stability, especially where livestock use is restricted (such as in CRP). While dense clubmoss provides soil stability on sites where it exists, anecdotal observations by some suggest that it competes for the limited water resources in the upper soil profile, which restricts plant available water; however, a study from Canada (Colberg and Romo 2003) in a similar climate on similar soils indicates that the correlation between reduced plant available water and clubmoss cover is negligible. Although quantitative evidence is currently unavailable, the correlation between reduced plant production and competition for space may simply be the cause. Dense patches of clubmoss may inhibit seed contact with the soil, reducing seedling recruitment.

Invasive weeds are beginning to have a high impact on this ecological site, particularly cheatgrass and field brome. These plants are typically associated with nearby human impacts (roads and construction activities).

#### Plant Communities and Transitions

A state and transition model for this ecological site is depicted below. Thorough descriptions of each state, transition, plant community, and pathway follow the model. This model is based on available experimental research, field data, field observations, and interpretations by experts. It is likely to change as knowledge increases.

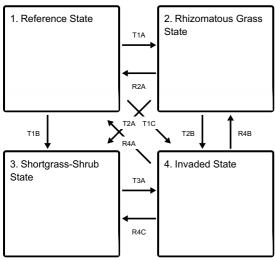
The plant communities within the same ecological site will differ across the MLRA due to the naturally occurring variability in weather, soils, and aspect. The biological processes on this site are complex; therefore, representative values are presented in a land management context. The species lists are representative and are not botanical descriptions of all species occurring, or potentially occurring, on this site. They are intended to cover the core species and the known range of conditions and responses.

Both percent species composition by weight and percent canopy cover are referenced in this document. Canopy cover drives the transitions between communities and states because of the influence of shade, the interception of rainfall, and the competition for available water. Species composition by dry weight remains an important descriptor of the herbaceous community and of the community as a whole. Woody species are included in the species composition for the site. Calculating the similarity index requires species composition by dry weight.

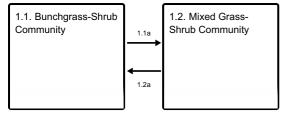
Although there is considerable qualitative experience supporting the pathways and transitions within the state and transition model (STM), no quantitative information exists that specifically identifies threshold parameters between grassland types and invaded types in this ecological site.

### State and transition model

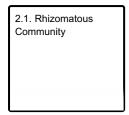
#### **Ecosystem states**



#### State 1 submodel, plant communities



#### State 2 submodel, plant communities



#### State 3 submodel, plant communities



#### State 4 submodel, plant communities



## State 1 Reference State

The Reference State consists of two Plant Communities. The first is dominated by deep-rooted bunchgrasses, shrubs, and rhizomatous grasses. The second is co-dominated by bunchgrasses and rhizomatous grasses. These communities regularly interchange in response to varying climatic conditions and minor disturbances. The Reference State exists where grazing by livestock and wildlife has historically been light to moderate and grazing events have been short. This reduces the opportunity for grazing regrowth and maintains.

## Community 1.1 Bunchgrass-Shrub Community

The Bunchgrass-Shrub Community is characterized by a plant community dominated by bluebunch wheatgrass, needle and thread, western wheatgrass, and Wyoming big sagebrush. This community is considered the Reference Community for this ecological site. Other species on this site are similar to others in this climate regime and include black greasewood, Gardner's saltbush, little sagebrush, scarlet globemallow, prairie Junegrass, and Sandberg bluegrass. Bluebunch wheatgrass lacks resistance to grazing during the critical growing season (spring) and will decline in vigor and production if grazed in the critical growing season more than one-year in three (Wilson et al. 1960). The reference state is moderately resilient and will return to dynamic equilibrium after a relatively short period of stress (such as drought or short-term improper grazing) if favorable or normal growing conditions and properly managed grazing are restored.

### **Dominant plant species**

- Wyoming big sagebrush (Artemisia tridentata ssp. wyomingensis), shrub
- greasewood (Sarcobatus vermiculatus), shrub

- winterfat (Krascheninnikovia lanata), shrub
- bluebunch wheatgrass (Pseudoroegneria spicata), grass
- green needlegrass (Nassella viridula), grass
- needle and thread (*Hesperostipa comata*), grass
- American vetch (*Vicia americana*), other herbaceous
- spiny phlox (*Phlox hoodii*), other herbaceous
- hairy false goldenaster (Heterotheca villosa), other herbaceous

### Table 5. Annual production by plant type

Plant Type	Low (Lb/Acre)	Representative Value (Lb/Acre)	High (Lb/Acre)
Grass/Grasslike	240	400	560
Shrub/Vine	45	75	105
Forb	15	25	35
Total	300	500	700

### Table 6. Ground cover

Tree foliar cover	0%
Shrub/vine/liana foliar cover	10-18%
Grass/grasslike foliar cover	45-65%
Forb foliar cover	5-10%
Non-vascular plants	0-1%
Biological crusts	0-3%
Litter	15-25%
Surface fragments >0.25" and <=3"	0-5%
Surface fragments >3"	0-5%
Bedrock	0%
Water	0%
Bare ground	15-25%

### Table 7. Soil surface cover

Tree basal cover	0%
Shrub/vine/liana basal cover	1-5%
Grass/grasslike basal cover	5-10%
Forb basal cover	1-3%
Non-vascular plants	0-1%
Biological crusts	0-3%
Litter	15-25%
Surface fragments >0.25" and <=3"	0-5%
Surface fragments >3"	0-5%
Bedrock	0%
Water	0%
Bare ground	15-30%

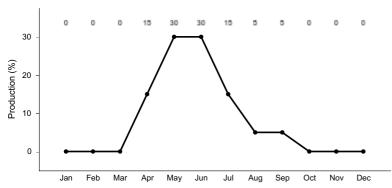


Figure 8. Plant community growth curve (percent production by month). MT44B032, Dry Uplands. Cool season grass dominated system. Most dry, upland sites located within MLRA 44B LRU A are characterized by early season growth which is mostly complete by Mid-July. Limited fall "greenup" if conditions allow..

## Community 1.2 Mixed Grass-Shrub Community

The Mixed Grass-Shrub Community is a community that contains nearly equal proportions of bluebunch wheatgrass and western wheatgrass as dominant species, with a slight increase in Wyoming big sagebrush. Western wheatgrass tolerates grazing pressure better than bluebunch wheatgrass and green needlegrass. The growing point for bluebunch wheatgrass grass is several inches above the ground, making it very susceptible to continued close grazing (Smoliack et al., 2006), while western wheatgrass growing points tend to be near the plant base. Western wheatgrass and Sandberg bluegrass increase in species composition when more palatable and less grazing-tolerant plants decrease due to improper grazing management. Western wheatgrass and bluebunch wheatgrass share dominance in the Mixed Grass-Shrub Community (1.2), with green needlegrass becoming rare. Other grass species that are more tolerant of grazing and are likely to increase in number compared to the Reference Plant Community include Sandberg bluegrass (Poa secunda), prairie Junegrass, and blue grama (Bouteloua gracilis). Hooded phlox (Phlox hoodii), scarlet globemallow (Sphaeralcea coccinea), hairy goldenaster (Heterotheca villosa), and pussytoes (Antennaria spp.) are examples of increaser forb species. Fringed sagewort (Artemisia frigida) is a shrub that also increases under prolonged drought or heavy grazing and can respond to precipitation that falls in July and August. Heavy, continuous grazing will reduce plant cover, litter, and mulch. The timing of grazing is important on this site because of the moisture limitations beyond June, especially on the drier sites. Bare ground will increase, exposing the soil to erosion. Litter and mulch will be reduced as plant cover declines. As long as bluebunch wheatgrass is still a dominant species in total biomass production, the site can return to the Bunchgrass-Shrub Community (Pathway 1.2A) under proper grazing management and favorable growing conditions. Western wheatgrass will continue to increase until it makes up the majority of the species composition. Once bluebunch wheatgrass has been reduced to less than 15 percent dry weight, it may be difficult for the site to recover to the Reference Plant Community (1.1). The risk of soil erosion increases when canopy cover decreases. As soil conditions degrade, there will be a loss of organic matter, reduced litter, and reduced soil fertility. Degraded soil conditions increase the difficulty of reestablishing bluebunch wheatgrass and returning to the Reference Community (1.1). The Mixed Grass-Shrub Community (1.2) is the at-risk plant community for this ecological site. When overgrazing continues, increaser species such as needle and thread and native forb species will become more dominant, and this triggers the change to the Rhizomatous State (2) or the Shortgrass-Shrub State (3). Until the Mixed Grass-Shrub Community (1.2) crosses the threshold into the Rhizomatous Community (2.1) or the Invaded Community (4.1), this community can be managed toward the Bunchgrass-Shrub Community (1.1) using prescribed grazing and strategic weed control. It may take several years to achieve this recovery, depending on growing conditions, the vigor of remnant bluebunch wheatgrass plants, and the aggressiveness of the weed treatments.

## Pathway 1.1a Community 1.1 to 1.2

Bluebunch wheatgrass loses vigor with improper grazing or extended drought. When vigor declines enough for plants to die or become smaller, species with higher grazing tolerance (in this ecological site, that would be both western and thickspike wheatgrass) increase in vigor and production as they access the resources previously used by bluebunch wheatgrass. The decrease in species composition by weight of bluebunch wheatgrass to less than 25

percent indicates that the plant community has shifted to the Mixed Bunchgrass Community (1.2). The driver for community shift 1.1A is improper grazing management or prolonged drought. This shift is triggered by the loss of vigor of bluebunch wheatgrass, soil erosion, or prolonged drought coupled with improper grazing. Blaisdell (1958) stated that drought and warmer-than normal temperatures are known to advance plant phenology by as much as one month. During drought years, plants may be especially sensitive or reach a critical stage of development earlier than expected.

## Pathway 1.2a Community 1.2 to 1.1

The Mixed Grass-Shrub Community (1.2) will return to the Bunchgrass-Shrub Community (1.1) with proper grazing management and appropriate grazing intensity. Favorable moisture conditions will facilitate or accelerate this transition. It will take several years of favorable conditions for the community to transition back to a bluebunch-dominated state. The driver for this community shift (1.2a) is the increased vigor of bluebunch wheatgrass, to the point that it represents more than 50 percent of species composition. The trigger for this shift is the change in grazing management favoring bluebunch wheatgrass. These triggers are generally conservative grazing management styles such as deferred or rest rotations utilizing light to moderate grazing (less than 50 percent use) combined with favorable growing conditions such as cool, wet springs. These systems tend to promote increases in soil organic matter, which promotes microfauna and can increase infiltration rates. Inversely, long periods of rest at a time when this state is considered stable may not result in an increase in bluebunch wheatgrass, and it has been suggested (Noy-Meir 1975) that these long periods of rest or underutilization may actually drive the system to a lower level of stability by creating large amounts of standing biomass, dead plant caudex centers, and gaps in the plant canopy.

## State 2 Rhizomatous Grass State

This state is characterized by having less than 10 percent bluebunch wheatgrass by dry weight. It is represented by one (1) community. Production in this state can be similar to that in the Reference State (1). Some native plants tend to increase under prolonged drought and heavy grazing practices. A few of these species may include western wheatgrass, needle and thread, Sandberg bluegrass, scarlet globemallow, hairy goldenaster, and fringed sagewort.

## Community 2.1 Rhizomatous Community

Long-term grazing mismanagement with continuous growing season pressure will reduce the total productivity of the site and lead to an increase in bare ground. Suppression of fire can also promote shrub growth, increasing plant interspaces. Once plant cover is reduced, the site is more susceptible to erosion and degradation of soil properties. Soil erosion or reduced soil fertility will result in reduced plant production. This soil erosion or loss of soil fertility indicates the transition to the Rhizomatous Grass State (2) because it creates a threshold requiring energy input to return to the Reference State (1). Transition to the Rhizomatous Grass Community (2.1) may be exacerbated by an extended drought. conditions. Western wheatgrass dominates the Rhizomatous Grass Community (2.1). Bluebunch wheatgrass makes up less than 10 percent of the species composition by dry weight, and the remaining bluebunch wheatgrass plants tend to be scattered and low in vigor. Invasive species will become more common. Hairy goldenaster, pussytoes, and scarlet globemallow are examples of increaser forbs. It is not uncommon for a minor component of invader species such as dandelion and yellow salsify (Tragopogon dubius) to be present, though at a level not high enough to be considered for transition to the Invaded State (4). This creates more competition for bluebunch wheatgrass and makes it difficult for bluebunch wheatgrass to quickly respond to a change in grazing management alone. Therefore, an input of energy is required for the community to return to the Reference State (1). Wind and water erosion may be eroding soil from the plant interspaces. Soil fertility is reduced, and soil surface erosion resistance has declined compared to the Reference State (1). This community crossed a threshold compared to the Mixed Grass-Shrub Community (1.2) due to the erosion of soil, vegetation composition, loss of soil fertility, or degradation of soil conditions. This results in a critical shift in the ecology of the site. The effects of soil erosion can alter the hydrology, soil chemistry, soil microorganisms, and soil structure to the point where intensive restoration is required to restore the site to another state or community. Changing grazing management alone cannot create sufficient improvement to restore the site within a reasonable time frame. Dormaar (1997) stated that with decreased grazing pressure, a needle and thread/blue grama plant community did not change species

composition, but the content of the soil carbon increased. It will require a considerable input of energy to move the site back to the Reference State (1). This state has lost soil or vegetation attributes to the point that recovery to the Reference State (1) will require reclamation efforts, i.e., soil rebuilding, intensive mechanical treatments, and reseeding. The transition to this state could result from overgrazing and fire suppression, especially repeated early-season grazing coupled with extensive drought. If heavy grazing continues, plant cover, litter, and mulch will continue to decrease, and bare ground will increase, exposing the soil to accelerated erosion. Litter and mulch will move off-site as plant cover declines. Continued improper grazing will drive the community into a Shortgrass-Shrub State (3). Introduction or expansion of invasive species will further drive the plant community into the Invaded State (4).

## State 3 Shortgrass-Shrub State

Shortgrass-Shrub State lacks midstatured bunchgrasses. Western wheatgrass, blue grama, Sandberg bluegrass, and prairie Junegrass are dominant grasses. Increaser shrubs like broom snakeweed and prickly pear cactus are almost as productive as larger sagebrush shrubs. Most shrub species are heavily hedged. This state will be treated as a terminal one (e.g., restoration will likely be impossible or unsuccessful and require major energy inputs).

## Community 3.1 Shortgrass-Shrub Community

In response to multiple, and often concurrent, disturbance events (overgrazing, continued drought, extreme fire, etc.), deep-rooted grasses, forbs, and shrubs are removed and replaced with shallow-rooted plants. The result is an unstable site that is highly susceptible to erosion as well as invasion by nonnative species, particularly annual grasses. Plant production is extremely limited and often too low to be economically valuable for grazing livestock.

## State 4 Invaded State

The Invaded State is identified as being in the exponential growth phase of invader abundance where control is a priority. Dominance (or relative dominance) of noxious or invasive species reduces species diversity, forage production, wildlife habitat, and site protection. A level of 10 percent invasive species composition by dry weight indicates that a substantial energy input will be required to create a shift to the grassland state (herbicide, mechanical treatment), even with a return to proper grazing management or favorable growing conditions. Prescriptive grazing can be used to manage invasive species. In some instances, carefully targeted grazing (sometimes in combination with other treatments) can reduce or maintain the species composition of invasive species.

## **Community 4.1 Annual Grass Invaded Community**

Communities in this state may be structurally indistinguishable from any of the communities in States 1 through 3, except annual grasses exceed 10 percent of species composition by dry weight. Typically, the Reference State is resistant to invasion from annual grasses; however, especially hot and dry conditions and a hot growing season may allow annuals to establish. For aggressive invasive species such as cheatgrass, early in the invasion process there is a lag phase where the invasive plant populations remain small and localized for long periods before expanding exponentially (Hobbs and Humphries 1995). Production in the invaded community may vary greatly. A site with early stages of invasion, where soil fertility and chemistry remain near reference, may have production near that of the reference community. A site with degraded soils and an infestation of cheatgrass may produce only 10 to 20 percent of the reference community. The Invaded State (4) is reached when invasive species dominate the site, either in terms of species composition by weight or in terms of their impact on the community. As annual bromes become established, they become very difficult to eradicate. Therefore, considerable effort should be put into preventing plant communities from crossing a threshold into the Invaded State (4) through early detection and proper management. Preventing new invasions is by far the most cost-effective control strategy, and typically places an emphasis on education. Control measures used on the noxious plant species impacting this ecological site include chemical, biological, and cultural control methods. The best success has been found with an integrated pest management (IPM) strategy that incorporates one or several of these options along with education and prevention

## Transition T1A State 1 to 2

The Reference State (1) transitions to the Rhizomatous Grass State (2) if bluebunch wheatgrass, by dry weight, decreases to below 15 percent or if bare ground cover is increased beyond 35 percent. The driver for this transition is loss of taller bunchgrasses, which creates open areas in the plant canopy with bare soil. Soil erosion results in decreased soil fertility, driving transitions to the Rhizomatous Grass State. There are several other key factors signaling the approach of transition T1A: increases in soil physical crusting, decreases in cover of cryptogamic crusts, decreases in soil surface aggregate stability and/or evidence of erosion including water flow patterns, development of plant pedestals, and litter movement. The trigger for this transition is improper grazing management and/or long-term drought leading to a decrease in bluebunch wheatgrass composition to less than 15 percent and reduction in total plant canopy cover.

## Transition T1B State 1 to 3

The Reference State (1) transitions to the Shortgrass-Shrub State (3) when Bluebunch wheatgrass is removed from the plant community and needle and thread is subdominant to short statured bunchgrasses such as Sandberg bluegrass. The trigger for this transition is loss of taller bunchgrasses, which creates open spaces with bare soil. Soil erosion results in decreased soil fertility, driving transitions to the Shortgrass-Shrub State. There are several other key factors signaling the approach of transition T1B: increases in soil physical crusting, decreases in cover of cryptogamic crusts, decreases in soil surface aggregate stability and/or evidence of erosion including water flow patterns, development of plant pedestals, and litter movement. The driver for this transition is improper grazing management, intense or repeated fires, and/or heavy human disturbance. Rapid transition is generally realized where livestock are confined to small pastures for long periods of time.

## Transition T1C State 1 to 4

Healthy plant communities are most resistant to invasion however, regardless of grazing management, without some form of active weed management (chemical, mechanical, or biological control) and without prevention the Reference State (1) can transition to the Invaded State (4) in the presence of aggressive invasive species such as spotted knapweed, leafy spurge, and cheatgrass. The Central Rocky Mountain Valleys tend to resists invasion of cheatgrass however repeated heavy grazing or intense human activities can open the interspaces of the bunchgrass community and allow for encroachment. Long-term stress conditions for native species (e.g., overgrazing, drought, and fire) accelerate this transition. If populations of invasive species reach critical levels, the site transitions to the Invaded State. The trigger for this transition is the presence of aggressive invasive species. Species composition by dry weight of invasive species approaches 10 percent.

## Restoration pathway R2A State 2 to 1

The Rhizomatous Grass State (2) has lost soil or vegetation attributes to the point that recovery to the Reference State (1) may require reclamation efforts such as soil rebuilding, intensive mechanical and cultural treatments, and/or revegetation. Examples of mechanical treatment may be brush control while cultural treatments may include prescribed grazing, targeted brush browsing, or prescribed burning. Prescribed grazing tactics may include a rest or deferred season rotation with light or moderate grazing intensity. Low intensity prescribed fires to reduce competitive increaser plants such as needle-and-thread and Sandberg bluegrass. A low intensity fire will also reduce Wyoming big sagebrush densities. In areas with potential of annual grass infestation, fire should be carefully planned or avoided. The drivers for this restoration pathway are reclamation efforts along with proper grazing management.

Transition T2A State 2 to 3

As improper grazing management continues vigor of bunch grasses will decrease, and the shorter grasses and shrubs will increase towards the Shortgrass-Shrub State (3). Prolonged drought will provide a competitive advantage to shrubs allowing them to become codominant with grasses. Shrub canopy will increase. Key transition factors: increase of native shrub canopy cover; reduction in bunchgrass production; decrease in total plant canopy cover and production; increases in mean bare patch size; increases in soil crusting; decreases in cover of cryptobiotic crusts; decreases in soil aggregate stability; and/or evidence of erosion including water flow patterns and litter movement.

## Transition T2B State 2 to 4

Invasive species can occupy the Rhizomatous Grass State (2) and drive it to the Invaded State (4). The Rhizomatous Grass State is at risk if invasive seeds and/or other viable material are present. The driver for this transition is more than 10 percent dry weight of invasive species. The trigger is the presence of seeds and/or other viable material of invasive species.

## Transition T3A State 3 to 4

Invasive species can occupy the Shortgrass-Shrub State (3) and drive it to the Invaded State (4). The Shortgrass-Shrub State is at risk of this transition occurring if invasive seeds or viable material are present. The driver for this transition is presence of critical population levels of invasive species. The trigger is the presence of seeds or viable material of invasive species. This state has sufficient bare ground that the transition could occur simply due to presence or introduction of invasive seeds or viable material. This is particularly true of aggressive invasive species such as spotted knapweed and cheatgrass. This transition could be assisted by overgrazing (failure to adjust stocking rate to declining forage production), long-term lack of fire, or extensive drought.

## Restoration pathway R4A State 4 to 1

Restoration of the Invaded State (4) to the Reference State (1) requires substantial energy input. The drivers for the restoration pathway are removal of invasive species, restoration of native bunchgrass species, persistent management of invasive species, and proper grazing management. Without continued control, invasive species are likely to return (probably rapidly) due to the presence of seeds and/or other viable material in the soil and management related practices that increase soil disturbance. If the site has transitioned from the Reference State to the Invaded State, the amount of time and resources necessary to transition back to the Reference may will be less than a site that transitioned from the Rhizomatous Grass State. Sites that transitioned from the Rhizomatous Grass State (2) will require more inputs to reestablish native bunchgrasses before it will resemble the Reference State (1). Sites that have transitioned from the Degraded State (3) to the Invaded State (4) may be severely lacking soil and vegetative properties that will allow for restoration to the Reference State. Hydrologic function damage may be irreversible especially with accelerated gully erosion.

## Restoration pathway R4B State 4 to 2

If invasive species are removed before remnant populations of bunchgrasses have been drastically reduced the Invaded State (4) can return to the Rhizomatous Grass State. The driver for the reclamation pathway is weed management without reseeding. Continued Integrated Pest Management (IPM) will be required as many of the invasive species that can occupy the Invaded State have extended dormant seed life. The trigger is invasive species control.

## Restoration pathway R4C State 4 to 3

If invasive species are removed the site could return to the Degraded State (3). Without sufficient remnant populations of preferred plants the Invaded State (4) is not likely to return to any of the other states. The driver for the reclamation pathway is weed management without reseeding. The trigger is invasive species control. The

invading species cause a significant increased soil loss due to lack of ground cover (Lacey et al. 1989).

## Additional community tables

Table 8. Community 1.1 plant community composition

Group	Common Name	Symbol	Scientific Name	Annual Production (Lb/Acre)	Foliar Cover (%)
Grass	/Grasslike	-		•	
1	Deep rooted bunchgra	asses	150–200		
	bluebunch wheatgrass	PSSP6	Pseudoroegneria spicata	125–150	25–40
	needle and thread	HECO26	Hesperostipa comata	50–75	5–10
	green needlegrass	NAVI4	Nassella viridula	10–65	5–10
	squirreltail	ELEL5	Elymus elymoides	0–20	0–1
2	Rhizomatous grasses	-		150–200	
	western wheatgrass	PASM	Pascopyrum smithii	130–180	5–10
	thickspike wheatgrass	ELLA3	Elymus lanceolatus	50–80	3–5
	plains reedgrass	CAMO	Calamagrostis montanensis	0–20	0–1
3	Increaser Bunchgrass	/Shortgras	ses	40–60	
	Sandberg bluegrass	POSE	Poa secunda	45–55	3–5
	prairie Junegrass	KOMA	Koeleria macrantha	20–40	0–5
	blue grama	BOGR2	Bouteloua gracilis	10–20	0–3
	saltgrass	DISP	Distichlis spicata	0–15	0–2
	needleleaf sedge	CADU6	Carex duriuscula	5–10	0–1
Shrub	/Vine			-	
4	Shrubs			60–90	
	Wyoming big sagebrush	ARTRW8	Artemisia tridentata ssp. wyomingensis	40–80	10–15
	greasewood	SAVE4	Sarcobatus vermiculatus	0–20	0–5
	little sagebrush	ARAR8	Artemisia arbuscula	0–20	0–5
	broom snakeweed	GUSA2	Gutierrezia sarothrae	0–10	0–1
	prairie sagewort	ARFR4	Artemisia frigida	0–10	0–1
	Gardner's saltbush	ATGA	Atriplex gardneri	0–5	0–1
	slender buckwheat	ERMI4	Eriogonum microthecum	0–5	0–1
	plains pricklypear	OPPO	Opuntia polyacantha	0–1	0
Forb	!		!		
5	Forbs			20–30	
	American vetch	VIAM	Vicia americana	10–20	1–3
	desertparsley	LOMAT	Lomatium	5–10	1–2
	scarlet globemallow	SPCO	Sphaeralcea coccinea	5–10	1–2
	bastard toadflax	соим	Comandra umbellata	5–10	1–2
	pussytoes	ANTEN	Antennaria	5–10	1–2
	spiny phlox	PHHO	Phlox hoodii	5–10	0–2
	milkvetch	ASTRA	Astragalus	0–10	0–2
	woolly plantain	PLPA2	Plantago patagonica	0–1	0–1

### **Animal community**

The Claypan ecological site provides a variety of wildlife habitats for an array of species. Prior to the settlement of this area, large herds of antelope, elk, and bison roamed. Though the bison have been replaced, mostly with domesticated livestock, elk and antelope still frequently utilize this largely intact landscape.

The relatively high grass component of the Reference Community provides excellent nesting cover for multiple neotropical migratory birds that select for open grasslands, such as the long-billed curlew and McCown's longspur.

Greater sage grouse may be present on sites with suitable habitat, typically requiring a minimum of 15 percent sagebrush canopy cover (Wallestad 1975). The Bunchgrass Community (1.1) will likely have the minimum sagebrush canopy cover present. The potential for a diverse forb component in the Reference State may provide important early-season (spring) foraging habitat for the greater sage grouse and their broods. Other communities on the site with sufficient sagebrush cover may harbor sage grouse populations, specifically Community 2.1 (Rhizomatous Grass-Shrub), where big sagebrush populations are under a reduced fire regime. Also, as the canopy cover of sagebrush increases in State 2 and, to a lesser extent, in Degraded State 3, pygmy rabbit, Brewer's sparrow, and mule deer use may increase.

Managed livestock grazing is suitable on this site; however, this site can be easily damaged by overgrazing due to its lower production and high bare ground percentage. In order to maintain the productivity of the site, grazing on adjoining sites with less production must be managed carefully to make sure utilization on this site is not excessive. Management objectives should include maintenance or improvement of the native plant community. Careful management of the timing and duration of grazing is important. Shorter grazing periods and adequate deferment during the growing season are recommended for plant maintenance, health, and recovery. According to McLean et al., early-season defoliation of bluebunch wheatgrass can result in high mortality and reduced vigor in plants. They also suggest, based on prior studies, that regrowth is necessary before dormancy to reduce injury of bluebunch wheatgrass.

The grazing season has a greater impact on winterfat than grazing intensity. Late winter or early spring grazing is detrimental. However, early winter grazing may actually be beneficial (Blaisdell 1984).

Continual non-prescribed grazing of this site will be detrimental, will alter the plant composition and production over time, and will result in the transition to the Altered State. The transition to other states will depend on the duration of poorly managed grazing as well as other circumstances such as weather conditions and fire frequency.

The Rhizomatous Grass State is subject to further degradation into the Shortgrass-Shrub State or Invaded State. Management should focus on grazing management strategies that will prevent further degradation, such as seasonal grazing deferment or winter grazing where feasible. Communities within this state are still stable and healthy under proper management. Forage quantity and quality may be substantially decreased from the Reference State.

Grazing is possible in the Invaded State. Invasive species are generally less palatable than native grasses. Forage production is typically greatly reduced in this state. Due to the aggressive nature of invasive species, sites in the Invaded State face an increased risk of further degradation. Grazing has to be carefully managed to avoid further soil loss and degradation and possible livestock health issues.

Prescriptive grazing can be used to manage invasive species. In some instances, carefully targeted grazing (sometimes in combination with other treatments) can reduce or maintain the species composition of invasive species. In the Shortgrass-Shrub State, grazing may be possible but is generally not economically or environmentally sustainable.

### **Hydrological functions**

The hydrologic cycle functions best in the Reference State (1) with good infiltration and deep percolation of rainfall; however, the cycle degrades as the vegetation community declines. Rapid rainfall infiltration, high soil organic matter, good soil structure, and good porosity accompany high bunchgrass canopy cover (Thurow et al. 1986). High ground cover reduces raindrop impact on the soil surface, which keeps erosion and sedimentation transport low. Water leaving the site will have a minimal sediment load, which allows for high water quality in associated streams.

High rates of infiltration will allow water to move below the rooting zone during periods of heavy rainfall. The Bunchgrass-Shrub Community (1.1) should have no rills or gullies present, and drainage ways should be vegetated and stable. Water flow patterns, if present, will be barely observable. Plant pedestals are essentially nonexistent. Plant litter remains in place and is not moved by wind or water.

In the Shortgrass Shrub State (3) and the Invaded State (4), canopy and ground cover are greatly reduced compared to the Reference State (1), which impedes the hydrologic cycle. Infiltration will decrease and runoff will increase due to reduced ground cover, the presence of shallow-rooted species, rainfall splash, soil capping, reduced organic matter, and poor structure. Sparse ground cover and decreased infiltration can combine to increase the frequency and severity of flooding within a watershed. Soil erosion is accelerated, the quality of surface runoff is poor, and sedimentation increases. (McCalla et al., 1984)

#### Recreational uses

This site provides some limited recreational opportunities for hiking, horseback riding, big game hunting, and upland bird hunting. Some plants have flowers that appeal to photographers. This site provides valuable open space.

## **Wood products**

none

### Other products

none

## Inventory data references

The information contained within this Ecological Site Description has been obtained from field observations, historic data, and professional judgement. Inventory sites are located across Southwest Montana.

### References

- . Fire Effects Information System. http://www.fs.fed.us/database/feis/.
- . 2021 (Date accessed). USDA PLANTS Database. http://plants.usda.gov.
- Arno, S.F. and G.E. Gruell. 1982. Fire History at the Forest-Grassland Ecotone in Southwestern Montana. Journal of Range Management 36:332–336.
- Barrett, H. 2007. Western Juniper Management: A Field Guide.
- Bestelmeyer, B., J.R. Brown, J.E. Herrick, D.A. Trujillo, and K.M. Havstad. 2004. Land Management in the American Southwest: a state-and-transition approach to ecosystem complexity. Environmental Management 34:38–51.
- Bestelmeyer, B. and J. Brown. 2005. State-and-Transition Models 101: A Fresh look at vegetation change.
- Blaisdell, J.P. 1958. Seasonal development and yield of native plants on the Upper Snake River Plains and their relation to certain climate factors.
- Blaisdell, J.P. and R.C. Holmgren. 1984. Managing Intermountain Rangelands--Salt-Desert Shrub Ranges. General Tech Report INT-163. USDA Forest Service, Intermountain Forest and Range Experiment Station, Ogden, UT.

- Bunting, S.C., B.M. Kilgore, and C.L. Bushey. 1987. Guidelines for Prescribe burning sagebrush-grass rangelands in the Northern Great Basin. General Technical Report INT-231. USDA Forest Service Intermountain Research Station, Ogden, UT. 33.
- Colberg, T.J. and J.T. Romo. 2003. Clubmoss effects on plant water status and standing crop. Journal of Range Management 56:489–495.
- Daubenmire, R. 1970. Steppe vegetation of Washington.
- DiTomaso, J.M. 2000. Invasive weeds in Rangelands: Species, Impacts, and Management. Weed Science 48:255–265.
- Dormaar, J.F., B.W. Adams, and W.D. Willms. 1997. Impacts of rotational grazing on mixed prairie soils and vegetation. Journal of Range Management 50:647–651.
- Hobbs, J.R. and S.E. Humphries. 1995. An integrated approach to the ecology and management of plant invasions. Conservation Biology 9:761–770.
- Kuchler, A.W. 1964. Potential natural vegetation of the conterminous United States.
- Lacey, J.R., C.B. Marlow, and J.R. Lane. 1989. Influence of Spotted knapweed (Centaurea maculosa) on surface runoff and sediment yield.. Weed Technology 3:627–630.
- Lesica, P. and S.V. Cooper. 1997. Presettlement vegetation of Southern Beaverhead County, MT.
- Manske, L.L. 1980. Habitat, phenology, and growth of selected sandhills range plants.
- Masters, R. and R. Sheley. 2001. Principles and practices for managing rangeland invasive plants. Journal of Range Management 38:21–26.
- McCalla, G.R., W.H. Blackburn, and L.B. Merrill. 1984. Effects of Livestock Grazing on Infiltration Rates of the Edwards Plateau of Texas. Journal of Range Management 37:265–269.
- McLean, A. and S. Wikeem. 1985. Influence of season and intensity of defoliation on bluebunch wheatgrass survival and vigor in southern British Columbia. Journal of Range Management 38:21–26.
- Miller, R.F., T.J. Svejcar, and J.A. Rose. 2000. Impacts of western juniper on plant community composition and structure. Journal of Range Management 53:574–585.
- Moulton, G.E. and T.W. Dunlay. 1988. The Journals of the Lewis and Clark Expedition. Pages in University of Nebraska Press.
- Mueggler, W.F. and W.L. Stewart. 1980. Grassland and Shrubland Habitat Types of Western Montana.

- Pelant, M., P. Shaver, D.A. Pyke, and J.E. Herrick. 2005. Interpreting Indicators of Rangeland Health.
- Pellant, M. and L. Reichert. 1984. Management and Rehabilitation of a burned winterfat community in Southwestern Idaho. Proceedings--Symposium on the biology of Atriplex and related Chenopods. 1983 May 2-6; Provo UT General Technical Report INT-172.. USDA Forest Service Intermountain Forest and Range Experiment Station. 281–285.
- Pitt, M.D. and B.M. Wikeem. 1990. Phenological patterns and adaptations in an Artemisia/Agropyron plant community. Journal of Range Management 43:350–357.
- Pokorny, M.L., R. Sheley, C.A. Zabinski, R. Engel, T.J. Svejcar, and J.J. Borkowski. 2005. Plant Functional Group Diversity as a Mechanism for Invasion Resistance.
- Ross, R.L., E.P. Murray, and J.G. Haigh. July 1973. Soil and Vegetation of Near-pristine sites in Montana.
- Schoeneberger, P.J. and D.A. Wysocki. 2017. Geomorphic Description System, Version 5.0..
- Smoliak, S., R.L. Ditterlin, J.D. Scheetz, L.K. Holzworth, J.R. Sims, L.E. Wiesner, D.E. Baldridge, and G.L. Tibke. 2006. Montana Interagency Plant Materials Handbook.
- Stavi, I. 2012. The potential use of biochar in reclaiming degraded rangelands. Journal of Environmental Planning and Management 55:1–9.
- Stringham, T.K., W.C. Kreuger, and P.L. Shaver. 2003. State and Transition Modeling: an ecological process approach. Journal of Range Management 56:106–113.
- Stringham, T.K. and W.C. Krueger. 2001. States, Transitions, and Thresholds: Further refinement fro rangeland applications.
- Sturm, J.J. 1954. A study of a relict area in Northern Montana. University of Wyoming, Laramie 37.
- Thurow, T.L., Blackburn W. H., and L.B. Merrill. 1986. Impacts of Livestock Grazing Systems on Watershed. Page in Rangelands: A Resource Under Siege: Proceedings of the Second International Rangeland Congress.
- Various NRCS Staff. 2013. National Range and Pasture Handbook.
- Walker, L.R. and S.D. Smith. 1997. Impacts of invasive plants on community and ecosystem properties. Pages 69–86 in Assessment and management of plant invasions. Springer, New York, NY.
- Wambolt, C. and G. Payne. 1986. An 18-Year Comparison of Control Methods for Wyoming Big Sagebrush in Southwestern Montana. Journal of Range Management 39:314–319.
- West, N.E. 1994. Effects of Fire on Salt-Desert shrub rangelands. Proceedings--Ecology and Management of Annual Rangelands: 1992 May 18-22. Boise ID General Technical Report INT-GTR-313.. USDA Forest Service Intermountain Research Station. 71–74.

Whitford, W.G., E.F. Aldon, D.W. Freckman, Y. Steinberger, and L.W. Parker. 1989. Effects of Organic Amendments on Soil Biota on a Degraded Rangeland. Journal of Range Management 41:56–60.

Wilson, A.M., G.A. Harris, and D.H. Gates. 1966. Cumulative Effects of Clipping on Yield of Bluebunch wheatgrass. Journal of Range Management 19:90–91.

### **Approval**

Kirt Walstad, 9/07/2023

### Rangeland health reference sheet

Interpreting Indicators of Rangeland Health is a qualitative assessment protocol used to determine ecosystem condition based on benchmark characteristics described in the Reference Sheet. A suite of 17 (or more) indicators are typically considered in an assessment. The ecological site(s) representative of an assessment location must be known prior to applying the protocol and must be verified based on soils and climate. Current plant community cannot be used to identify the ecological site.

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Date	05/10/2020
Approved by	Kirt Walstad
Approval date	
Composition (Indicators 10 and 12) based on	Annual Production

### **Indicators**

the reference condition.

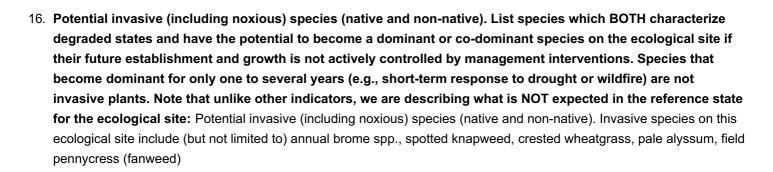
1.	Number and extent of rills: Rills are not present in the reference condition.
2.	Presence of water flow patterns: Water flow patterns are rarely present except after heavy rainfall events. If present, flow patterns will be short (less than 2 feet in length).
3.	Number and height of erosional pedestals or terracettes: Pedestals are not evident in the reference condition.
4.	Bare ground from Ecological Site Description or other studies (rock, litter, lichen, moss, plant canopy are not bare ground): Bare ground is 20-25%. It consists of small, randomly scattered patches.

5. Number of gullies and erosion associated with gullies: Gullies are not present in the reference condition.

6. Extent of wind scoured, blowouts and/or depositional areas: Wind scoured, or depositional areas are not evident in

7.	Amount of litter movement (describe size and distance expected to travel): Litter movement is not evident in the reference condition.
8.	Soil surface (top few mm) resistance to erosion (stability values are averages - most sites will show a range of values): The average soil stability rating is 4-5 under plant canopies and 3 in plant interspaces. The A horizon thickness is highly variable and is 1-4 inches thick.
9.	Soil surface structure and SOM content (include type of structure and A-horizon color and thickness): Soil Structure at the surface is medium granular. A Horizon should be 1-4 inches thick with color, when wet, typically ranging in Value of 4 or less and Chroma of 3 or less.  Local geology may affect color, it is important to reference the Official Series Description (OSD) for characteristic range. https://soilseries.sc.egov.usda.gov/osdname.aspx
10.	Effect of community phase composition (relative proportion of different functional groups) and spatial distribution on infiltration and runoff: Infiltration of the Thin Clay Pan ecological site is slow to very slow. The site is well drained. An even distribution of mid stature bunchgrasses (40 percent), rhizomatous grass (35 percent), cool season shortgrasses (10-15 percent), forbs (1-5 percent), and shrubs (5-15 percent) optimizes infiltration and reduces runoff under normal moisture events.
11.	Presence and thickness of compaction layer (usually none; describe soil profile features which may be mistaken for compaction on this site): A compaction layer is not present in the reference condition. Soil profile will contain an abrupt transition to an argillic horizon which can be misinterpreted as compaction, however, the soil structure will be strong medium subangular blocky, where a compaction layer will be platy or structureless (massive).
12.	Functional/Structural Groups (list in order of descending dominance by above-ground annual-production or live foliar cover using symbols: >>, >, = to indicate much greater than, greater than, and equal to):
	Dominant: Mid-statured, cool season, perennial bunchgrasses (Bluebunch wheatgrass) = rhizomatous grasses
	Sub-dominant: shrubs > shortgrass grasses/grasslikes (blue grama, Junegrass) > forbs
	Other:
	Additional:
13.	Amount of plant mortality and decadence (include which functional groups are expected to show mortality or decadence): Mortality in herbaceous species is not evident. Species with bunch growth forms may have some natural mortality in centers is 3 percent or less.
14.	Average percent litter cover (%) and depth ( in): Total litter cover ranges from 25-30%. Most litter is irregularly distributed on the soil surface and is not at a measurable depth.

15.	Expected annual annual-production (this is TOTAL above-ground annual-production, not just forage annual-
	production): Average annual production is 500lbs/acre. Low: 300lbs/acre High 700lbs/acre. Production varies based on
	effective precipitation and natural variability of soil properties for this ecological site.
	Metric: Low: 336 kg/ha Average (RV): 560 kg/ha High: 785 kg/ha



Native species such as broom snakeweed, Sandberg's bluegrass, blue grama, pricklypear cactus, greasewood, etc. when their populations are significant enough to affect ecological function, indicate site condition departure.

17. **Perennial plant reproductive capability:** In the reference condition, all plants are vigorous enough for reproduction either by seed or rhizomes in order to balance natural mortality with species recruitment.